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# Ambient air pollution, social inequalities and asthma exacerbation in Greater Strasbourg (France) metropolitan area: The PAISA study

D. Bard¹, O. Laurent¹,², S. Havard¹,³, S. Deguen¹, G. Pedrono⁴, L. Filleul⁵,
C. Segala⁴, A. Lefranc⁶, C. Schillingerⁿ and E. Rivièreⁿ
¹Département d'Epidémiologie et de Recherche Clinique, École des Hautes Études en Santé
Publique, Rennes
²LEPID, Institut de Radioprotection et de Sûreté Nucléaire, Fontenay aux Roses
³INSERM U 707, Paris
⁴SEPIA-santé, Baud
⁵CIRE Aquitaine, Institut national de Veille Sanitaire, Bordeaux
⁶DSE, Institut national de Veille Sanitaire, Saint Maurice
¹Association pour la Surveillance et l'Etude de laPollution Atmosphérique en Alsace,
Schiltigheim

#### 1. Introduction

The socio-economic status (SES) of populations has an influence on the incidence or mortality rates of numerous health outcomes, among which respiratory diseases (Prescott et al., 2003; Ellison-Loschmann et al., 2007). Considering asthma, the possible contribution of SES to overall prevalence -regardless of asthma severity-, remains controversial in industrialized countries. Several studies indicate that allergic asthma is more prevalent in more well-off populations whereas the non-allergic forms of asthma are more common in the deprived ones (Cesaroni et al., 2003; Blanc et al., 2006). On the other hand, severe asthma whatever its etiology appears to be more frequent in the latter populations, as compared to the more affluent (Basagana et al., 2004). Risk factors for exacerbations (e.g., passive smoking (Wright Subramanian, 2007), psychosocial stress (Gold & Wright, 2005), cockroach allergens (Kitch et al., 2000), and suboptimal compliance with anti-inflammatory medication (Gottlieb et al., 1995)) are generally more common among people with asthma and low SES than their better-off counterparts. These observations support the hypothesis that some factors more present in deprived populations contribute to asthma exacerbation (Mielck et al., 1996). One candidate factor is ambient air pollution, a well-demonstrated risk factor for asthma (Samet & Krewski, 2007). Epidemiologic studies report short-term associations (latency of several hours to several days) between exposure to air pollutants (particulate matter, nitrogen dioxide [NO<sub>2</sub>], sulfur dioxide, and ozone [O<sub>3</sub>]) and various asthma attack indicators routinely available: hospitalization (Sunyer et al., 1997), emergency department

(ED) visits (Romeo et al., 2006), calls to mobile emergency medical service networks (Medina

et al., 1997) and visits to doctors' offices for asthma attacks (Hajat et al., 1999). Associations are also reported with indicators measured directly in individuals with asthma: respiratory symptoms and consumption of short-acting  $\beta$ -agonist (SABA) drugs (von Klot et al., 2002; Schildcrout et al., 2006).

It should be stressed that in such studies, be they time series (Sunyer et al., 1997), case-crossover (Jalaludin et al., 2008), or panel (Ward & Ayres, 2004; Romeo et al., 2006) studies, investigators average citywide ambient pollutant concentrations for estimating exposure, although these concentrations often vary spatially and strongly within cities (Jerrett et al., 2005a), with few exceptions (Tolbert et al., 2000; Neidell, 2004; Erbas et al., 2005). Failure to consider these spatial variations at the sub-city level can lead to exposure misclassification and resulting bias (Jerrett et al., 2005a). Conversely, taking these spatial variations into account in studying the relations between air pollution and health would allow more rigorous measurement of the magnitude of these relations (Jerrett et al., 2005a). Another problem inherent in ecologic studies is spatial autocorrelation, which expresses the non-independence of geographic observations (Haynes et al., 2001). With very rare exceptions (Jerrett et al., 2001; Buzzelli et al., 2003), studies of environmental equity do not consider this phenomenon, although it is essential to avoid violating the hypotheses that underlie the application of statistical models, and may thus lead to erroneous conclusions about associations.

Another key issue is identification of the populations most susceptible to air pollution and the precise measurement of exposure-response relations in these subsets (Maynard et al., 2003; Samet & Krewski, 2007). Several studies show that some specific populations, such as children, the elderly, and people with chronic diseases, including people with diabetes and chronic obstructive pulmonary disease, are more susceptible to air pollution than the general population (Annesi-Maesano et al., 2003; Samet & Krewski, 2007). Some suspect that socioeconomic deprivation influences the relations between air pollution and health (O'Neill et al., 2003). Many investigators have tested this hypothesis with regard to mortality, and their overall results suggest the existence of larger relative risks for more deprived populations (Laurent et al., 2007). Substantially fewer researchers have tested this hypothesis with regard to asthma attacks (Norris et al., 1999; Lin et al., 2004; Neidell, 2004), and the few that have done so have considered socioeconomic indicators measured at resolutions ranging from the individual level (Neidell, 2004; Kim et al., 2007) to the level of geographic residence areas of various sizes (Nauenberg & Basu, 1999; Norris et al., 1999; Lin et al., 2004; Son et al., 2006; Kim et al., 2007). Whether or not a socioeconomic indicator is an effect modifier, however, may well depend on the resolution at which it is measured (Laurent et al., 2007). To our knowledge, only two studies have tested the influence of socioeconomic indicators measured at the level of small areas (Canadian enumeration areas (Lin et al., 2004) and US zip codes (Neidell, 2004) on the short-term relations between air pollution and asthma attacks. Disadvantaged people may be more susceptible to asthma attacks when they face additional environmental insults, such as ambient air pollution (Maynard et al., 2003). Nonetheless, the few studies (Nauenberg & Basu, 1999; Annesi-Maesano et al., 2003; Maynard et al., 2003; O'Neill et al., 2003; Lin et al., 2004; Laurent et al., 2007) that have tested this hypothesis have produced divergent results. This may be because the distribution according to SES of factors modulating the relations between air pollution and asthma morbidity differs between the study settings or because of lack of statistical power.

In the absence of individual data, which are not generally routinely available, ecological (or contextual) measures of SES are frequently used to describe health inequalities. Although some epidemiological studies are based on only one socioeconomic indicator (income, educational level, or occupation), e.g. (Kunst et al., 1998; Finkelstein et al., 2003), SES is usually recognized as complex and multidimensional, integrating different components that may be either material (e.g., housing conditions, income, or occupation), social (e.g., social position or isolation, or family support) or both (Folwell, 1995; Braveman et al., 2005). Area-based deprivation indices for the measurement of the economic or social disadvantages of urban areas were proposed in the 1980s (Townsend, 1987). Initially designed for health care planning and resource allocation, they have been recently used to evaluate and analyse health inequalities (Carstairs, 1995; Niggebrugge et al., 2005; Eibner Sturm, 2006). These measures, including Townsend's, Carstairs', and Jarman's indices, as well as the more recent Index of Multiple Deprivation, combine contextual indicators such as unemployment rate or proportions of overcrowded or of non-owner-occupied households (Jarman, 1983; Townsend et al., 1988; Carstairs & Morris, 1991; DETR, 2000). Since the end of the 1990s, numerous other area-based deprivation indices have emerged in the United States (Singh, 2003; Eibner & Sturm, 2006; Messer et al., 2006), Canada (Pampalon & Raymond, 2000), New Zealand (Salmond et al., 1998), Japan (Fukuda et al., 2007), Italy (Tello et al., 2005), Spain (Benach & Yasui, 1999), and Belgium (Lorant, 2000). Hospitalizations or emergency department visits reflect severe asthma exacerbation episodes. As to increase sensitivity, alternative indicators integrating less severe events may be used, such as telephone calls to medical services or asthma exacerbation specific drugs. Naureckas et al. (Naureckas et al., 2005) showed that SABA sales are valuable indicators of asthma morbidity and its temporal variability. These sales are especially good predictors of the risk of emergency department visits and hospitalization for asthma attacks in the days immediately afterwards (Naureckas et al., 2005). Moreover, they generally reflect less severe morbidity states than those requiring emergency department visits or hospitalizations and are therefore very sensitive indicators. Finally, they are very common, likely to involve large numbers of individuals, even within small geographic units and short observation periods (Pitard et al., 2004; Naureckas et al., 2005). While SABA drugs are the treatment of choice for asthma attacks (Global Initiative for Asthma, 2007), they are also prescribed for acute exacerbation of chronic obstructive pulmonary disorders (COPD) (Urbano & Pascual, 2005). Nevertheless, this disease is rare in people less than 40 years old (Viegi et al., 2007); within this age group, SABA prescriptions and sales are therefore very specific for asthma attacks. The PAISA study objectives were to investigate the short-term relations between air pollution estimated at the level of small geographic areas and asthma attacks and to test the influence of socioeconomic deprivation, also measured at this level, on these relations. Two different, although related, indicators were tested, that is i) calls for emergency services for asthma exacerbation and ii) SABA drugs deliveries.

#### 2. Methods

#### 2.1 Setting and statistical unit

The setting of our study was the Strasbourg metropolitan area (SMA), gathering municipalities (316 km²) with a population of about 450,000 inhabitants located in the Bas-Rhin district (or 'département', an administrative subdivision) in the Alsace region of North-

Eastern France. The statistical small-area unit used was the French census residential block (Ilots Regroupés pour l'Information Statistique-IRIS), a submunicipal division devised by the National Institute for Statistics and Economic Studies (INSEE). This unit is the smallest geographic area in France for which socioeconomic and demographic information from the national census is available. It is, in terms of population size, intermediate between US census tracts (about 4,000 inhabitants) and US census block groups (about 1,000 inhabitants) with an average number of inhabitants of 2,000. The division of neighborhoods into census blocks takes into account the physical obstacles that may break up urban landscapes (important traffic arteries, bodies of water, green spaces, etc.) and aims to maximize their homogeneity in population size, socioeconomic characteristics, land use and zoning. The SMA is subdivided into 190 blocks, with areas ranging from 0.05 km² to 19.6 km² (median values of 0.45 km² for the entire SMA and 0.29 km² for the city center). Fifteen blocks covering a very small population (250 people, altogether 0.2 percent of the total population) were removed from the study because of confidentiality rules in force in France.

#### 2.2 Socioeconomic Status Indicator

The demographic and socioeconomic data come from the 1999 national population census, conducted by INSEE. The census database is structured into different domains, including employment, family and household, educational level, housing, immigration status, and income; together these regroup a collection of very diverse quantitative variables, such as proportions of unemployed people, foreigners, blue-collar workers, households without cars, people aged 15 years or older with general or vocational maturity certificates, etc. Accordingly, we selected 52 variables from the available data, endeavoring to apply theoretical concepts of deprivation (Krieger et al., 1997), relying on the indicators most often used in the literature (Morris & Carstairs, 1991; Pampalon & Raymond, 2000; Challier & Viel, 2001; Jordan et al., 2004), and using the same definitions as INSEE most often uses in its studies. All but 2 of the 52 variables were proportions (the exceptions were mean number of people per room and median income per consumption unit). Some variables, intentionally redundant, were introduced into the analysis to determine statistically the most discriminating for characterizing deprivation (*e.g.*, unemployment among total labor force, among men, among women).

*Constructing the index* 

We developed the socioeconomic deprivation index by principal component analysis (PCA) of the 52 variables. All variables were first standardized (*i.e.* centered and reduced) to remove the influence of different units of measure and thus to give them the same weight in the analysis. This statistical technique permits the synthesis of information contained in a great number of variables by constructing new and independent synthetic variables – the principal components. These new variables are linear combinations of some of the initial variables, defined by a maximum variance. The principal advantage of this method is that it makes it possible to consider relations between the variables, by attributing to each a weight that takes the relations into account (that is, treating them as coefficients of the linear combinations).

In our study, to facilitate the use of the deprivation index in a public health context, we chose to construct a single index for all the blocks to maximizing the variance of the first principal component. Several consecutive PCA were thus performed to conserve only the variables most strongly correlated with the first component and contributing most to its

construction. If two redundant variables were conserved, only the one most closely correlated with the first component was finally retained (and in each case, that turned out to be the variable not stratified by sex). Finally, this first component, used as our socioeconomic deprivation index, was calculated for each block, as the linear combination of the final variables retained.

To obtain the most discriminant cartographic representation of socioeconomic disparities, the blocks were grouped into five classes of index values, approximately equivalent to the quintiles (Beguin & Pumain, 1994). The first class comprised the least deprived blocks and the fifth the most deprived blocks. The choropleth mapping of the index was performed with the Geographic Information System ArcView<sup>TM</sup> v. 9.1 (ESRI Inc., Redlands, CA, USA).

#### Index validity

According to Coste et al. (1995) (Coste et al., 1995), several criteria make it possible to judge the validity of composite measurement scales (in our case, a socioeconomic deprivation index), in particular, its content validity and its construct validity (Fermanian, 1996).

Content validity is assessed by the extent to which the items composing the scale or index are relevant to and representative of all of the possible items that can describe the phenomenon measured. To select the initial subset of variables for analysis, we carefully ensured that each census domain (income, job, housing, family and household, etc.) was represented by at least one variable, and again at the end we verified this representativeness for all of the final variables.

Construct validity verifies that from a theoretical point of view the instrument is associated with the concept it is supposed to measure. It tests the relations between the index variables, both internally (factorial validity and internal consistency) and externally (convergent validity), and thus identifies more precisely the real meaning of the concept measured by the indicator. The internal validity of our index was estimated with the results of the PCA and Cronbach's alpha coefficient. This coefficient judges how the variables retained by the PCA measure a one-dimensional concept - here, socioeconomic deprivation. The closer the coefficient value is to one, the better the verification that the index variables are homogenous. The convergent validity was tested using Pearson's correlation coefficients with the most widely used British indices (those of Carstairs and Townsend) (Townsend et al., 1988; Carstairs & Morris, 1991). These indices combine four variables, three common to both: (1) proportion of unemployed, (2) proportion of households without a car, and (3) proportion of over- crowded households. Townsend's index also includes the concept of rental versus owner-occupied housing (proportion of non-owner-occupied housing) while Carstairs' index considers a variable related to the household head's place in the British social scale (proportion of households in which the householder belongs to social classes 4 or 5). These indices were compiled at the census block level by adapting or translating some British variables when they had no direct correspondence in the French census database (for example, the variable "proportion of households in which the householder belongs to social class 4 or 5" was translated for the French census database as "the proportion of blue-collar workers").

Index reliability

To test the robustness of our index, the same methodological procedure as that used for the SMA (the same 52 initial variables, same algorithm decision for the final choice of variables, the same tests of validity) was conducted for another French metropolitan area at the

resolution of the census block – the Lille metropolitan area (LMA). The LMA was selected because its demographic and socioeconomic characteristics are quite distinct from those of the SMA (*e.g.*, 489 census blocks for a total population around 1,100,000 inhabitants). The objective of this stage was to demonstrate that similar results could be obtained in a different part of France and to show that the variables selected for the construction of the index were independent of the study area and characterized socioeconomic level in as discriminating a manner as possible.

#### **Application**

Census blocks were divided into five deprivation categories according to their index value; the first category comprised the most privileged blocks and the fifth the most deprived ones. More details can be found in (Havard et al., 2008).

#### Calls to emergency medical services for asthma exacerbation

The two mobile emergency and healthcare networks operating in the SMA (Service d'Aide Médicale d'Urgence (SAMU) and SOS Médecins provided data about emergency telephone calls made to physicians for asthma attacks. The Bas-Rhin emergency medical services unit (SAMU 67) is a public service that coordinates pre- hospital emergency medical services. SOS Médecins of Strasbourg is a private service providing emergency general medical care. We included each call regarding an asthma attack that reached either SAMU or SOS Médecins from January 1, 2000, to December 31, 2005. The two databases were merged, duplicates being excluded. Each call was geo-coded to the census block where the patient was located at the time of the call, based on the postal address. Only 2 percent of the calls could not be geocoded; they were excluded from the analysis.

#### **SABA** sales

The data about SABA sales to people living in the SMA during 2004 came from four health insurance funds: the regional union of health insurance funds (Union Régionale des Caisses d'Assurance Maladie [URCAM], for salaried workers and very low-income families), agricultural social insurance fund (Mutualité Sociale Agricole [MSA], for farmers and similar workers), the fund for self-employed workers (Régime Social des Indépendants [RSI]), and the Lorraine students' insurance fund (Mutuelle Générale des Etudiants de Lorraine [MGEL]). These funds cover >90% of the local population (Com-Ruelle et al.). In 2004, these funds either paid or reimbursed (fully or partially) all SABA sales prescribed by a physician. All records of SABA sales in pharmacies were extracted from the databases of these four insurance funds by requests for code R3A4 of the European Pharmaceutical Market Research Association nomenclature (Tollier et al., 2005). The following information was furnished for each sale: date, age group (0 to 9, 10 to 19, 20 to 39 years), sex, and census block of residence of the person for which SABA were prescribed. All sales were combined into a single file for analysis.

#### Ambient air pollution

Hourly ambient concentrations of particulate matter less than  $10~\mu m$  in aerodynamic diameter (PM<sub>10</sub>), nitrogen dioxide, sulfur dioxide, and ozone were modeled by the local air quality monitoring association (Association pour la Surveillance et l'Etude de la Pollution Atmosphérique en Alsace, Schiltigheim) for each block during the entire study period (Bard et al., 2007). Modeling was conducted with Atmospheric Dispersion Modeling System (ADMS-Urban), a deterministic Gaussian dispersion model (McHugh et al., 1997) that integrates emissions inventories, meteorological data (wind direction and speed, temperature and cloudiness, supplied by Météo France, the French meteorological service), and background

pollution measurements as input parameters. A limitation of the model was the estimation of Monin-Obukhov length from cloudiness measurements, but the correlation coefficients for the modeled and effectively measured ambient concentrations were 0.73 for  $PM_{10}$ , 0.87 for nitrogen dioxide, 0.84 for ozone. However, this correlation was only 0.06 for sulfur dioxide. Therefore, sulfur dioxide was no longer considered in the analyses.

#### Potential confounders

Daily meteorological variables (temperature, atmospheric pressure, and relative humidity) were obtained from Météo France, and daily pollen counts were obtained from the National Network of Aerobiological Surveillance (Thibaudon, 2004). Weekly influenza case counts came from the Sentinelles network (Flahault et al., 2006) of the National Institute of Health and Medical Research.

#### Statistical analyses

As some census blocks featured a relatively small population size, yielding unstable rates, we carried out an empirical Bayesian smoothing using STIS<sup>TM</sup> software (Space Time Intelligence System 2007 (Ann Arbor, MI, USA: Terraseer), except for drug sales, where smoothing was not warranted in view of the much greater number of events per census block (Marshall, 1991). Positive and statistically significant spatial autocorrelations were found for our SES index (Moran's I = 0.54, p<0.01), rates and standardized incidence ratios (SIR) for asthma calls (I from 0.67 to 0.77 according to age group, p<0.01). Associations between SES and health event rates or SIR were quantified by Pearson's correlation test and their significance assessed using a modified Student's t-test as proposed by Clifford et al., (Clifford et al., 1989) as to accounting for spatial autocorrelation of neighboring census blocks variables.

Associations between asthma calls and air pollution were assessed with case-crossover models (Maclure & Mittelman, 1991), which are similar to case-control models, except that in the former, the subjects serve as both cases and controls, depending on when they are considered. The subject serves as a case on the day of the health event and a control on days without any health event. Control days were defined according to a monthly time-stratified design (Janes et al., 2005). For an asthma call occurring on a given weekday (e.g., Monday), control days were chosen as the same days of the week throughout the rest of the month (thus, three or four days; here, the other Mondays of the month). Conditional logistic regression was employed for analyses. The statistical significance of the odds ratios was tested by means of a two-sided  $\chi^2$  test at the 5 percent level.

Base models

We first analyzed all calls or SABA sale, without differentiating them according to socioeconomic deprivation. Associations between health events and ambient air pollution concentrations modeled by census block were estimated, adjusting for holidays, meteorological variables (daily maximum temperature, maximum atmospheric pressure, and mean relative humidity), influenza epidemics, and pollen counts.

We tested the influence of air pollution indicators averaged on the day of the call (lag 0) and then averaged on the day of the call and the 1–5 previous days (lag 0–1 to 0–5) or up to 10 days in the case of SABA sales. The daily air pollution indicator considered for  $PM_{10}$ , nitrogen dioxide, and sulfur dioxide was the 24-hour average concentration, and for ozone it was the maximum daily value of the 8-hour moving average. The analysis for ozone considered only health events occurring between April 1 and September 30 of each year, because of the very low concentrations of this pollutant in winter.

These associations were assessed for cases of all ages and then for groups of cases aged 0–19, 20–64, and >64 years, except for SABA sales (below 40). These groups were defined according to the guidelines of the French Data Protection Authority, to ensure the confidentiality of the health data geocoded by census blocks.

Testing the influence of deprivation

Interactions with socioeconomic deprivation were tested for the lag times for which the strongest associations were observed in the basic models.

Deprivation was introduced first as a discrete variable. The SMA population was divided into five strata with contrasting deprivation levels, according to the quintiles of the distribution of the deprivation index. An odds ratio for the relation between asthma calls and pollution was estimated for each stratum. The heterogeneity of the odds ratios between these strata was assessed by means of a two-sided  $\chi^2$  test at the 5 percent level (Atkinson et al., 2001). These analyses were conducted for the same age groups as in the base models.

In an alternative method, deprivation was introduced as a continuous variable. For that purpose, a case-crossover model was constructed for each of the 174 census blocks in which asthma calls occurred during the study period. The models did not converge in the census blocks with fewer than 11 asthma calls. These census blocks were treated as follows. Those adjoining neighboring census blocks of a similar deprivation level (differences of deprivation index values less than one 10th of the deprivation index scale) were merged with them. The deprivation index value attributed to each newly created geographic unit was the mean of the deprivation index values from the merged blocks, weighted by their respective population counts. Census blocks (n = 26) that bordered no neighboring census block of a similar deprivation level (differences of deprivation index values one 10th or more of the deprivation index scale) were excluded from these analyses.

The heterogeneity of the odds ratios estimated in the 136 geographic units so defined was assessed as described above (Atkinson et al., 2001). The influence of the deprivation level of the geographic units on the odds ratios was tested by linear regression (weighted by the inverse of the variance of these odds ratios) in both fixed- and random-effects models. These analyses were conducted for  $PM_{10}$ , nitrogen dioxide, and sulfur dioxide for cases of all ages but were not feasible for ozone (April–September) and specific age groups, because of the small numbers of calls in most census blocks.

All analyses were carried out using SAS statistical software v 9.1 and 8.2 (SAS Institute, Cary, NC, USA).

#### 3. Results

#### 3.1 Socioeconomic Status Indicator

Our deprivation index was defined by the PCA as the first principal component, which explained 66% of the total variance of the model, while the second explained only 17%. The impossibility of drawing any residual structure from the second component justifies the conclusion that all of the discriminating information useful for characterizing socioeconomic deprivation was in fact captured by the first component. Our index is composed of 19 socioeconomic variables that describe different dimensions of socioeconomic deprivation. Seventeen can be qualified as primarily material variables and only two directly refer to social dimension of the deprivation (that is, single-parent family and foreign population). Although several of the variables retained belong to the same domain, we consider that they

provide complementary information for characterizing the deprivation. Our index opposes the concepts of material and, to a lesser extent, social deprivation (it is significantly and positively correlated with variables relative to unemployment, immigration, social isolation, household overcrowding, low educational level, and low income) to those of well-being and economic comfort (it is significantly and negatively correlated with variables relative to job stability, high educational level, better living conditions, and high income). Moreover, these variables had differing relative influences in the construction of the index (*e.g.*, proportion of unemployed people contributed to 6.72% of the variance and the percentage of households without cars to only 3.91%).

Table I presents the mean values of the index and of the 19 variables by deprivation class. The first class (C1) is characterized by the lowest mean index value and includes the most privileged blocks, defined by the best living conditions (low percentages of the variables describing socioeconomic deprivation and high percentages of variables illustrating material comfort). At the other end, the fifth class (C5) is characterized by the highest mean index value and includes the most deprived blocks. The blocks of this class are characterized by more socioeconomic deprivation, a greater cumulative lack of material and social resources (the highest percentages of variables describing deprivation and the lowest percentages of variables related to positive living conditions). For example, on average, only 1.28% of homes in the C1 blocks were low income subsidized housing while this percentage exceeded 75% for the C5 blocks.



Table I. Deprivation Index and mean values of its 19 variables, according to deprivation categories<sup>a</sup>

Variables	Census domain		Depriv	ation categ	ories	
		C1	C2	C3	C4	C5
Blue collar workers in the labor force	Job	17.40	15.80	17.20	21.70	36.40
Primary residences that are houses or farms	Housing	61.40	28.60	7.10	5.80	3.60
Primary residences that are multiple dwelling units	Housing	37.10	69.50	90.20	90.90	95.10
Households without a car	Housing	8.80	16.00	27.60	35.80	36.50
Households with≥2 cars	Housing	48.90	32.40	18.90	14.90	14.00
People with insecure jobs in the labor force	Job	8.30	10.40	13.70	16.90	17.40
People aged ≥15 yrs with general or vocational maturity certificates	Education	13.30	12.20	11.90	9.50	6.60
People aged ≥15 yrs with at least a lower postsecondary education	Education	10.90	10.30	10.30	8.10	3.50
People aged >15 yrs with no more than a completed elementary education	Education	9.80	9.80	10.80	16.20	33.50
Non owner-occupied primary residences	Housing	28.20	48.90	72.10	80.30	91.30
People with stable jobs in the labor force	Job	76.40	72.00	68.20	63.10	54.90
Subsidized housing among all primary residences	Housing	1.30	6.80	9.30	27.30	75.60
Single-parent families	Families and households	7.70	10.30	13.20	15.70	23.60
Median income/consumption unit (in Euros/yr) <sup>c</sup>	Income	19,977	19,819	17,120	14,270	8,402
Primary residences with more than 1 person/room	Housing	3.20	4.30	5.90	7.90	21.40
Foreigners in the population	Immigration status	3.20	6.20	7.90	12.30	22.0
Mean no. people/room <sup>c</sup>	Housing	0.60	0.60	0.64	0.69	0.86
Unemployed people in the labor force	Job	5.30	7.50	10.20	13.50	24.20
People unemployed >1 yr in the labor force	Job	2.40	3.20	4.40	5.80	11.70
Deprivation index		1.09	1.75	2.25	2.67	3.44

<sup>&</sup>lt;sup>a</sup>All variables are percentages (%), unless otherwise stated; p value for trend < 0.01 for all variables. <sup>b</sup>C1 is the least deprived category, C5 is the most. <sup>c</sup>Not a percentage.

relatively high population and comprising the urban municipalities.

Conversely, nearly 50% of the households in C1 blocks had two cars or more compared with 14% for the C5 blocks (Table I).

Our index thus demonstrates the existence of a socio-economic gradient in the SMA, along which we observe a progressive degradation of level and quality of life between the first and fifth classes (p < 0.01). The different variables used to construct the index sometimes vary significantly in the same direction as the gradient, and sometimes in the opposite direction (Table I), depending on the aspect (positive or negative) of SES that they express (p < 0.01). This gradient appears clearly in the cartographic representation (Figure 1). The most affluent blocks are located on the outskirts of the study area and constitute a peri-urban ring around Strasbourg. These are principally blocks with a relatively low population, comprising rural municipalities. Conversely, socioeconomic deprivation is accentuated as we approach Strasbourg and reaches its maximum at the centre of the metropolitan area. The most disadvantaged blocks are principally those in Strasbourg and the inner suburbs with a

Validity tests

Content validity. The 19 variables retained cover all of the census domains and thus insure that our index has fully integrated the multiple socioeconomic aspects of deprivation (Table I)

Construct validity. The high Cronbach's  $\alpha$  coefficient (0.92) supports the hypothesis that these variables measure a one-dimensional concept – socioeconomic deprivation. Convergent validity is verified by very high correlation coefficients with the indices of both Townsend (0.97; p<0.01) and Carstairs (0.96; p < 0.01). *Reliability* 

Following the same process, 20 variables were retained for Lille metropolitan area, 18 of them were the same as for SMA and they explained the same proportion of variance (66%). Only three variables differed between the two indices: the percentages of the self-employed and of foreign immigrants were retained only for the Lille index and the percentage of blue-collar workers only for Strasbourg. Mapping the index for LMA again illustrates the strong spatial heterogeneity of SES between urban centers and outer suburbs. Finally, as for Strasbourg, all the tests of the index conclusively demonstrated its validity (*i.e.*, the Lille index had a high Cronbach's  $\alpha$  coefficient and was very highly correlated with both British indices; data not shown).

#### Calls to emergency medical services for asthma exacerbation

There were 4,682 calls for asthma (SAMU and SOS Médecins combined) and 1,173 SAMU-confirmed asthma cases (from 1,578 calls given to SAMU - some but not all initially reported to be for asthma). Cases finally diagnosed as asthma by participating physicians were treated as such in our analyses, regardless of the initial reason for call. We observed no trend in the annual number of health events.

Figure 2 shows the results of empirical Bayesian smoothing of calls for asthma exacerbation SIRs.

Maps of ambient air pollutant concentrations averaged by census block across the SMA (for the period 2000–2005) appear on figure 3.

As expected for a small-area study, high and statistically significant spatial autocorrelation was found for the deprivation index (I = 0.54, p < 0.01) and for health event indicators (Table II). These are mapped respectively in figures 4 and 5.

The age-specific rates and SIRs of calls for asthma increased from the least towards the most deprived blocks (Figure 4). Elevated ( $R \ge 0.65$ , p < 0.01) correlations were observed between the deprivation index and both the SIRs and rates of calls for the groups aged 20 years or older (Table II). Correlations were lower but still statistically significant for the younger groups.

Results were similar for SAMU-confirmed asthma cases (Table II). Correlations with the deprivation index were elevated ( $R \approx 0.62\text{-}0.63$ , p < 0.01) for those aged 20 years or older, and lower (but still significant) for the younger groups.

Table III presents the distribution of asthma calls and air pollutant concentrations for the entire SMA and for the five socioeconomic deprivation strata. Overall, there were 4,677 usable asthma calls made during the study period. As we previously reported, the numbers of calls increased from the least deprived stratum to the most deprived, despite their similar population denominators.

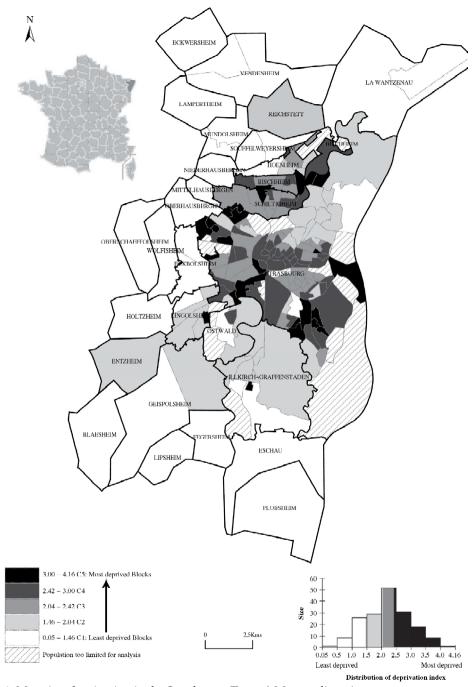


Fig. 1. Mapping deprivation in the Strasbourg (France) Metropolitan Area.

Table II. Summary of health data for the Strasbourg metropolitan area and their correlations with deprivation index

			All c	All calls for asthma	ıa			A	sthma case.	Asthma cases confirmed by SAMU	y SAMU	
Age	number	Mean number	% of	Annual rate per	Moran's	Correlation with	number	Mean % of all number number SAMU-	% of all SAMU-	Annual rate per	Moran's	Moran's Correlation with
group	of calls	group of calls of calls calls per block	calls	1,000 inhabitants	index	index deprivation of cases of cases confirmed index per block cases inl	of cases	of cases c per block	confirmed cases	1,000 nabitants	index	index
6-0	539	3.1	11.5	1.760	829.0	0.52†	135	8.0	11.5	0.437	0.71 §	0.26†
10-19	416	2.4	8.8	1.208	0.74s	0.47+	120	0.7	10.2	0.349	0.68 §	0.20
20-39	951	5.4	20.3	1.037	0.758	0.65+	249	1.4	21.2	0.272	0.64 §	0.63†
40-64	1138	6.5	24.3	1.464	0.70s	0.70+	377	2.11	32.1	0.485	0.71 §	0.62†
e5 +	1637	9.3	35	4.589	0.758	0.68†	292	1.7	24.9	0.819	99.0	0.63†
Total*	otal* 4682	26.8	100	1.752	0.77s	0.77†	1173	6.7	100	0.442	0.738	1990
* Mora	n's index	and corre	elation	Moran's index and correlations with SES index calculated for the age-standardised incidence ratio (SIR)	ndex calc	ulated for th	e age-star	ndardised	incidence	ratio (SIR)		

sp<0.01, estimated by a standard t-test †p<0.01 estimated by Clifford, Richardson and Hémon's test †p<0.05 estimated by Clifford, Richardson and Hémon's test

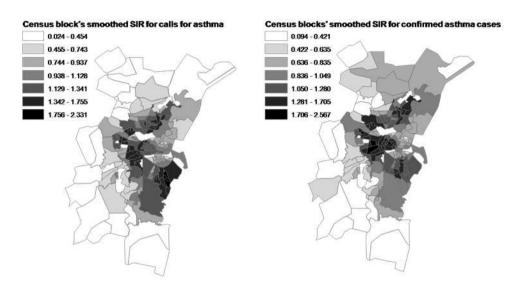


Fig. 2. Maps of age-standardized incidence ratios of emergency calls for asthma and of SAMU-confirmed asthma cases, across Strasbourg Metropolitan Area's census blocks

Table III. Distribution of population, asthma calls and air pollutant concentrations (Strasbourg Metropolitan Area, 2000-2005)

Deprivation stratum †	Population	Calls, all ages	Calls, age 0- 19	Calls, age 19-64	Calls, age over 64	PM <sub>10</sub>	SO <sub>2</sub>	NO <sub>2</sub>	O <sub>3</sub> §
Stratum 1	80,917	313	70	125	118	21	7.6	30.2	63.4
Stratum 2	92,534	784	159	300	325	22.1	8.6	35	58.6
Stratum 3	93,392	959	169	384	406	23.2	9.6	39	55
Stratum 4	105,367	1,243	226	574	443	23	9.3	38.3	55.5
Stratum 5	74 695	1,378	330	703	345	22.5	8.8	35.7	58
Overall	446,905	4,677	954	2,086	1,637	22.6	8.9	36.0	57.7

<sup>†</sup> Stratum 1 is the least deprived, and stratum 5 the most deprived

 $<sup>^{\</sup>ddagger}$  PM<sub>10</sub>, particulate matter less than 10 micrometers in aerodynamic diameter; NO<sub>2</sub>, nitrogen dioxide; SO<sub>2</sub>, sulfur dioxide. Concentrations for these three pollutants are averaged for period 2000-2005, in microgram per cubic meter

 $<sup>\</sup>S$  O<sub>3</sub>, ozone. Numbers reported are maximum 8-hour daily concentrations, averaged for summer months (April 1 to September 30) of period 2000-2005, in microgram per cubic meter

Base models

Figure 5 shows the odds ratios for the relations between asthma calls (all ages) and pollutant concentrations averaged on the day of the call (lag 0) and the previous 1–5 days (lag 0–1 to lag 0–5). The highest positive associations were observed for lag 0–1 for  $PM_{10}$  (for a 10- $\mu$ g x m<sup>-3</sup> increase, odds ratio (OR) = 1.035, 95 percent confidence interval (CI): 0.997, 1.075) and sulfur dioxide (OR = 1.056, 95 percent CI: 0.979, 1.139) and for lag 0 for nitrogen dioxide (OR = 1.025, 95 percent CI: 0.990, 1.062). The association for  $PM_{10}$  approached statistical significance (p= 0.07). For ozone, no association was observed (OR = 0.998, 95 percent CI: 0.965, 1.032).

Table IV presents the results of analyses by age group for these same lags. For  $PM_{10}$  (lag 0–1), associations were stronger for people younger than age 20 years and older than age 64 years as compared with people of all ages. For nitrogen dioxide (lag 0), higher odds ratios were observed in people over age 64 years than in people of all ages. For ozone, positive associations were observed only in persons older than 64 years.

Testing the influence of deprivation

Deprivation as a discrete variable. Figure 6 presents the odds ratios for the five deprivation strata among people of all ages. The odds ratios show no clear trend according to deprivation level for any pollutant. We detected no heterogeneity of the odds ratios between strata. No clear trends by deprivation were observed in specific age groups (Figure 7).

Deprivation as a continuous variable. Significant heterogeneity (p < 0.05) was detected between the odds ratios estimated for the 136 geographic units for each of the three pollutants considered ( $PM_{10}$ , and nitrogen dioxide). Heterogeneity remained significant after we discarded the highest and lowest 5 percent of the odds ratios in the distribution. Negative linear regression coefficients (table V) suggested that the values of the odds ratios for these pollutants tended to decrease slightly from the least deprived geographic units to the most deprived (Figure 8). The associations between the odds ratios and deprivation were not significant, however, regardless of the pollutant or the model (fixed or random effects).

#### **SABA** sales

Table VI presents the distribution of SABA sales and ambient pollutant concentrations for the entire SMA and for the five socioeconomically different strata. In 2004, SABA drugs were dispensed on 15,121 separate occasions for subjects aged 0 to 39 years. The number of sales per inhabitant slightly increased from the least to the most deprived stratum(Bard et al., 2007). The mean ambient pollutant concentrations over the study period were  $20.8\mu g/m^3$  for PM<sub>10</sub> (range, 1.2 to  $106.3~\mu g/m^3$ ),  $35.0~\mu g/m^3$  for nitrogen dioxide (range, 6.4 to 84.3  $\mu g/m^3$ ), and  $58.7\mu g/m^3$  for ozone (range, 2.6 to  $220.0~\mu g/m^3$ ).

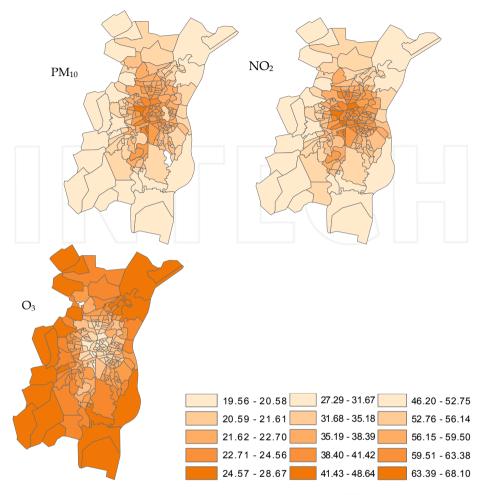


Fig. 3. Maps of ambient air pollutant concentrations (μg/m³) averaged by census block across the Strasbourg Metropolitan Area (2000–2005)

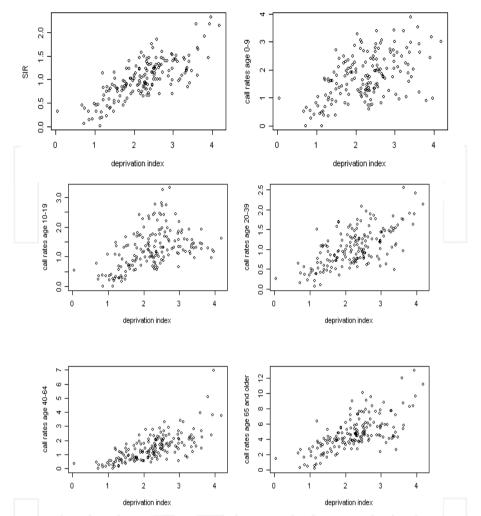


Fig. 4. Scatterplots of the rates of emergency calls for asthma according to deprivation of census blocks, for different age groups (on the left, the most privileged census blocks, on the right the most disadvantaged ones). SIR: Standardized Incidence Ratio

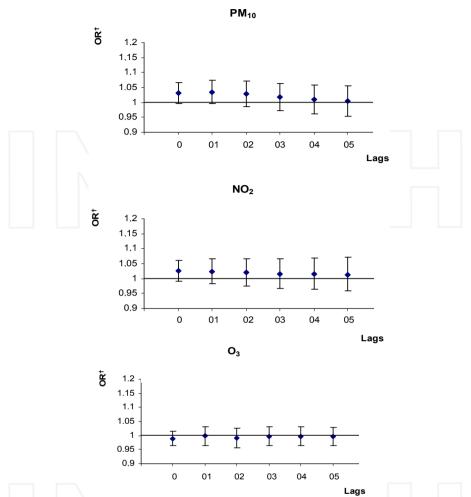


Fig. 5. Odds ratios between asthma calls and pollutant concentrations\*, for various lag times (Strasbourg Metropolitan Area, 2000-2005)

<sup>\*</sup> PM<sub>10</sub>, particulate matter less than 10 micrometers in aerodynamic diameter; NO<sub>2</sub>, nitrogen dioxide; O<sub>3</sub>, ozone.

 $<sup>\</sup>dagger$  Odds ratios reported for a 10  $\mu$ g.m<sup>-3</sup> increase in pollutant concentrations, adjusted for temperature, relative humidity, atmospheric pressure, holidays, influenza epidemics and pollen counts. Lag 0 is for pollutant concentrations averaged on the day of the call, lag 01 for pollutant concentrations averaged on the day of the call and the previous one, and so on.

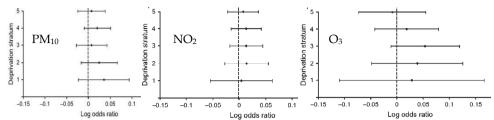


Fig. 6. Log odds ratio for emergency asthma calls per 10  $\mu$ g/m³ increase in ambient pollutant concentrations, according to stratum of socioeconomic deprivation, Strasbourg, France, 2000–2005. Stratum 1 was the least deprived, and stratum 5 was the most deprived. PM<sub>10</sub>: particulate matter less than 10  $\mu$ m in aerodynamic diameter; NO<sub>2</sub>: nitrogen dioxide; O<sub>3</sub>: ozone. Bars, 95% confidence interval.

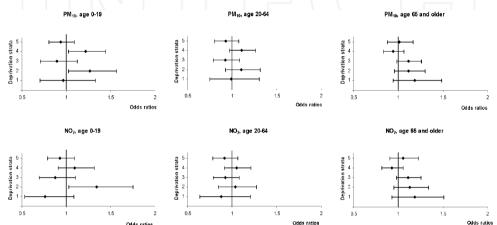


Fig. 7. Odds ratios between asthma calls and pollutants\*, for five strata of deprivation levels† and different age groups (Strasbourg Metropolitan Area, 2000-2005)

- \*  $PM_{10}$ , particulate matter less than 10 micrometers in aerodynamic diameter;  $NO_2$ , nitrogen dioxide;  $O_3$ , ozone. Odds ratios reported for a  $10 \, \mu g/m^3$  increase in pollutant concentrations, adjusted for temperature, relative humidity, atmospheric pressure, holidays, influenza epidemics and pollen counts
- † Stratum 1 is the least deprived, and stratum 5 is the most

#### Base model

Figure 9 presents the odds ratios for the individual lags for subjects aged 0 to 39 years. For all three pollutants, odds ratios greater than one were observed for lags from 4 to 10 days. For  $PM_{10}$  and nitrogen dioxide, most of the odds ratios were statistically significant (p<0.05) for lags from 5 to 10 days. For a lag of 2 days, odds ratios significantly lower than one were observed for all three pollutants. The structure of the lags reported in Figure 9 did not change when the analyses were stratified by smaller age subgroups (0 to 19 years, 20 to 39 years). No association was observed between SABA sales and the mean pollutant concentrations of the day of the sale and the 1 to 10 preceding days (lags 0 to 1 to 0 to 10).

On the basis of the associations reported in Figure 9, the lags kept for the subsequent analyses were lag 4 to 7 (mean concentrations of days 4, 5, 6, and 7 for  $PM_{10}$ , lag 4 to 10 for nitrogen dioxide, and lag 4 to 6 for ozone). A  $10~\mu g/m^3$  increase in ambient concentrations of  $PM_{10}$ , nitrogen dioxide, and ozone for these lags was associated with increases of 7.5% (95% CI, 4 to 11.2%), 8.4% (95% CI, 5 to 11.9%), and 1 (95% CI, - 0.3 to 2.2%), respectively, in SABA sales.

Testing the influence of deprivation

Figure 10 presents the odds ratios for the five different socioeconomic strata for the optimal lags mentioned above for all subjects aged 0 to 39 years. We observed no trend toward an increase or decrease of these odds ratios according to SES or when we stratified the analyses by smaller age subgroups (0 to 19 years, 20 to 39 years).

#### 4. Discussion

#### 4.1 Socioeconomic Status Indicator

From available census variables, we developed and validated a new French socioeconomic deprivation index at the census block level. Working at this resolution ensures a better homogeneity of the residents' characteristics. As a result, it is possible to better identify and measure with greater precision their relation with the health events observed. Available French census data combines material and, to a lesser extent, social aspects of deprivation, with no indication of a possible association with the health events to which it may be associated. In this study, it allowed to identify a socioeconomic gradient in the utilization of mobile emergency medical services for asthma attacks in the SMA (Laurent et al., 2008). This scale has also been shown to be robust (validity criterion) and reproducible for another French metropolitan area with different socioeconomic characteristics.



Table IV. Odds ratios between asthma calls and pollutant concentrations for different age groups and for selected lag times (Strasbourg Metropolitan Area, 2000-2005)

		\						\	\			
		All ages			Age 0-19			Age 20-64		/	Age 64 +	
Lag	3 Odds	%26	# <i>d</i>	Odds	%26	<b>#</b> <i>d</i>	Odds	%26	# <i>d</i>	Odds	%26	$b^*$
++	‡ ratios	0		ratios	confidence		ratios	confidence		ratios	confidence	
	ss			ss	intervals		œ	intervals		œ	intervals	
DM.* 0.1	1 025	0.997,	0.07	1 047	0.961,	000	1 020	0.964,	070	1 040	0.985,	717
5	1.033	1.075	0.0	1.04/	1.141	0.29	1.020	1.079	0.43	1.049	1.116	0.14
	1001	0.990,	0.16	1 000	0.926,	700	1 01 5	0.964,	D 11	1 050	0.660	010
NO2: 0	1.023	1.062	0.10	C00.1	1.086	0.94	1.013	1.070	0.57	1.050	1.113	0.10
* 00	7 0 1	0.979,	7	7	0.945,	010	0000	0.882,	0 0	7	0.975,	7
T-0		1.139	0.13	1.122	1.334	0.19	0.990	1.112	0.07	1.102	1.245	0.17
-	0000	0.965,	0	7700	0.891,	070	0000	0.949,	0	5	0.960,	011
		1.032	0.90	0.900	1.048	0.40	166.0	1.047	0.50	C10.1	1.074	60.0

' P $M_{10}$ , particulate matter less than 10 micrometers in aerodynamic diameter;  $NO_2$ , nitrogen dioxide.

<sup>‡</sup>Lag 0 is for pollutant concentrations averaged on the day of the call, lag 01 for pollutant concentrations averaged on the day of the tO<sub>3</sub>, ozone. Study period extending from April 1 to September 30 of each year

§ Odds ratios reported for a 10 µg.m.³ increase in pollutant concentrations, adjusted for temperature, relative humidity, atmospheric call and the previous one.

pressure, holidays, influenza epidemics and pollen counts

# Estimated by a two-sided  $\chi^2$  test

Table V. Influence of deprivation introduced as a continuous variable on the relations between pollutants and asthma calls (Strasbourg Metropolitan Area, 2000-2005)

		Dependent varia	ables: odds ratios be and asthma calls	tween pollutants
		$NO_2^{\dagger}$	SO₂ <sup>†</sup>	$PM_{10}^{\dagger}$
Independent variable:	Fixed effects model	-0.00272, p=0.49 ‡	-0.01032, p=0.18 ‡	-0.00244, p= 0.48 ‡
deprivation index	Random effects model	-0.00284, p=0.48 ‡	-0.01052, p=0.18 ‡	-0.00249, p=0.47 ‡

<sup>†</sup> Odds ratios reported for a 10  $\mu$ g.m<sup>-3</sup> increase in pollutant concentrations, adjusted for temperature, relative humidity, atmospheric pressure, holidays, influenza epidemics and pollen counts. PM<sub>10</sub>, particulate matter less than 10 micrometers in aerodynamic diameter; NO<sub>2</sub>, nitrogen dioxide.

Nonetheless, the strength of the correlations observed with the British indices by Townsend and Carstairs (correlations greater than 0.95 for both) suggest that we might have simply used them in our study, rather than constructing our own index. Indeed, their indices are more straightforward to use than ours (unweighted sum of standardized variables) and require fewer variables for their construction (4 compared with 19 in ours). However, they rely exclusively on purely material indicators, thus missing the social dimension of deprivation. The direct use of the British indices in a French setting is also limited by several factors. Firstly, these indices use data from the British census that do not correspond directly to French census items and for which an adaptation or translation is required (for example, the variable "proportion of households in which the householder belongs to social class 4 or 5" considered by Carstairs must be translated as "the proportion of blue-collar workers" for use in the French census database). Secondly, they were developed at a specific geographical level (enumeration district or ward) that is not equivalent to the French census block in terms of population size and socioeconomic characteristics. The effects of theses adaptations (i.e., variable definitions and geographical units) on the associations observed with health events have never, as far as we know, been assessed and do not encourage their use in our context. A last limitation of these indices is their inability to take into account differences between rural and urban areas because they include in their composition only variables specific to urban conditions (Gilthorpe & Wilson, 2003). The use in our study of the Index of Multiple Deprivation, constructed especially to overcome this weakness (Jordan et al., 2004) should therefore be a useful option. The French census data do not, however, lend themselves to the use of this index, for its composition includes 33 variables, many of which were either unavailable or could not be transposed from the census data. Our index thus appears to be a good alternative to the existing British indices: it is simpler to implement than the Index of Multiple Deprivation (19 compared with 33 variables, no weighting, etc.) and at the same time it overcomes the principal limitations of Carstairs and Townsend's indices, that is, it is validated at the resolution of census blocks and able to contrast deprived urban and advantaged rural areas in our study. This study nonetheless has its own

<sup>‡</sup> Beta coefficients for regression of the values of the odds ratios on the deprivation index. Statistical significance was estimated by a two-sided  $\chi^2$  test

limitations, the first of which is the need to acknowledge that one size does not fit all (Braveman et al., 2005) Our index is a comprehensive and composite measure of SES and was constructed independently of the health events to which it might be associated. It includes variables related to education, income, occupation, etc. and in that respect is a better indicator than a single variable taken in isolation to describe SES as a whole.

Characterizing in more detail the social component of deprivation, both at the individual level and accounting for neighborhood influences would allow shedding more light on the complex and intricate relationships between socioeconomic characteristics and the health impact of exposure to environmental stressors. Nonetheless, in the absence of individual data, ecological studies of this type remain necessary and useful for a better understanding of the interactions between socioeconomic factors and health.

#### Calls to emergency medical services for asthma exacerbation

Emergency calls for asthma attacks were positively, although not significantly, associated with concentrations of  $PM_{10}$ , and nitrogen dioxide modeled by census blocks. No association was observed for ozone. Overall, associations were higher among people younger than 20 years and older than 64 years. Socioeconomic deprivation measured by census block did not appear to influence these relations.

The daily exposure estimates we used were modeled for small areas and are geographically more precise than those usually employed to study short-term relations between air pollution and health. Compared with ambient concentrations averaged citywide, our exposure estimate likely reduced ecological biases (Jerrett et al., 2005b). Practically, the exposure measurement attributed to each subject was the concentration estimated for the census block where each patient was when the emergency network was called. However, we do not actually know whether the patients were in the same block in the hours to days preceding the call, and this fact obviously determines the extent to which our exposure measurement adequately reflects subjects' true exposure.

This point matters mainly for subjects who are frequently away from the neighborhood they live in -principally in the 20-64 year-old age group, globally characterized by mobility and autonomy, and who work for a living (often outside the neighborhood of residence). Conversely, in general the elderly rarely commute out of their neighborhoods of residence and when they do so, they generally go shorter distances (Benlahrech et al., 2001). Children also have more limited mobility than people aged 20-64 years (Agence de l'Environnement et de la Maîtrise de l'Energie/Institut de Radioprotection et de Sûreté Nucléaire, 2003) and generally attend the schools closest to their homes.

These points support the idea that our exposure measurement is more accurate for people aged younger than 20 and older than 64 years. These subjects are more likely to have called the emergency networks from their neighborhood of residence (and thus be geocoded in it). Moreover, for these subjects, the measurement of air pollution in the neighborhood of residence is more likely to provide an adequate reflection of exposure integrated over the days preceding the call than for more mobile subjects.

The ranges of the associations we found for the base models were similar to those reported by other studies of emergency calls (Medina et al., 1997) and visits to hospital emergency departments (Galan et al., 2003) for asthma. Above all, the associations observed for  $PM_{10}$  were very close to those reported by two meta-analyses of the associations between this pollutant and asthma symptoms (Weinmayr et al., 2010; Kunzli et al., 2000).

Small-area deprivation (introduced either as a discrete or as continuous variable) did not appear to influence the relations between ambient air pollution and asthma attacks. Of the six previous studies that investigated the influence of socioeconomic indicators on these relations (Nauenberg & Basu, 1999; Norris et al., 1999; Lin et al., 2004; Neidell, 2004; Son et al., 2006; Kim et al., 2007), five reported higher relative risks for populations with less advantageous socioeconomic characteristics (Nauenberg & Basu, 1999; Lin et al., 2004; Neidell, 2004; Lee et al., 2006; Kim et al., 2007). However, one of these studies found evidence of interaction according to the ecological socioeconomic indicator considered, and no interaction with the individual indicator (Kim et al., 2007). The sixth study reported slightly higher relative risks for the most deprived populations (Norris et al., 1999). Nevertheless, formal comparison of the results of these studies is difficult, as they focused on socioeconomic indicators measured at very heterogeneous resolutions (Laurent et al., 2007).

Three of these studies focused on socioeconomic indicators measured at very coarse geographic resolutions. Lee et al. (Lee et al., 2006) and Kim et al. (Kim et al., 2007) focused on Gu neighborhoods (of around 400,000 inhabitants) in the city of Seoul (Korea), for children aged 0-15 (Lee et al., 2006) and people of all ages, respectively (Kim et al., 2007). Overall, these studies reported higher relative risks between  $PM_{10}$ ,  $SO_2$ , and  $NO_2$  concentrations and asthma hospitalizations in the most deprived Gu. Norris et al., in Seattle (Washington, USA) (Norris et al., 1999) studied associations between the same pollutants and visits to emergency departments for asthma in children aged 0-18. They reported slightly lower relative risks for residents of the inner city, that is, the most deprived areas.

Two other studies focused on individual socioeconomic indicators in people of all ages. Kim et al (Lee et al., 2006), in Seoul (Korea) found that associations between  $PM_{10}$  and visits to emergency departments for asthma did not vary according to the annual amount of taxes paid to the national health insurance system. In contrast, Nauenberg and Basu, in Los Angeles (USA), found that effects of  $PM_{10}$  were greater in people with a less favorable health insurance status (Nauenberg & Basu, 1999).

Last, two studies focused on socioeconomic indicators measured by small areas. Neidell (Neidell, 2004) observed that carbon monoxide and ozone had a greater effect on asthma hospitalizations of Californian children aged 3-18 in ZIP codes characterized by lower educational attainment. This study, however, estimated associations between asthma hospitalizations and air pollutants on the basis of monthly indicators, which are inadequate to study short-term relations between these factors. Lin et al studied children aged 6-12 years in Vancouver (Canada) (Lin et al., 2004) and reported higher associations between nitrogen dioxide, sulfur dioxide and hospitalization for asthma in enumeration areas with lower household income levels.

The study by Lin et al (Lin et al., 2004) is the most comparable in design to ours, but reports somewhat different results. The reasons for this are unclear. Pollutant concentrations were very similar in the two settings. Although the studies used different types of exposure measurements (pollutant concentrations averaged citywide for Lin et al (Lin et al., 2004) and modeled by census blocks in ours), this difference does not explain the variation in findings. Indeed, alternative analysis in the SMA with citywide average exposure measurements did not noticeably change our results about interactions with deprivation. Lack of statistical power appears to be a plausible explanation for the difference between our results and those of Lin et al. For comparable age groups (0-20 versus 6-12) we had one quarter the number of health events to analyze.

Another point is that findings of interaction with deprivation are not necessarily transposable from one setting to another. If small-area deprivation does exert an influence on the relations between air pollution and asthma attacks, it would most likely be mediated through "third" factors that in some (but perhaps not in all) settings would be distributed unequally according to deprivation. Previous studies report that several factors thought to strengthen the associations between air pollution and asthma attacks are more present in deprived than in well-off neighborhoods. Among these are the prevalence of (both active and passive) smoking (Diez-Roux et al., 2003), psychosocial stress (Gold & Wright, 2005), unhealthy eating habits (Diez-Roux et al., 1999), amounts of indoor allergens (Kitch et al., 2000), inadequate compliance with anti-inflammatory medication (Gottlieb et al., 1995) and (a plausible result of the factors mentioned above) a higher ratio of severe to moderate forms of asthma among subjects with asthma (Basagana et al., 2004). Nevertheless, the distribution of these factors according to small-area deprivation may differ between study settings, due, for instance, to differences in climate (affecting allergen proliferation), social and cultural characteristics of local populations (influencing, among other things, eating and smoking habits) or effectiveness of health systems (influencing prescription of and compliance with anti-inflammatory medication).

Moreover, although we observed no interactions with small-area deprivation in the SMA, this does not rule out the existence here of interactions by socioeconomic factors measured at other resolutions (individual, household, geographic areas more or less fine than the French census block), as the study by Kim et al clearly illustrates (Kim et al., 2007). The use of multilevel models, which make it possible to assess more precisely the influence of factors (e.g. socioeconomic characteristics) measured at different resolutions, would be useful in studying this question further (O'Neill et al., 2003).

#### SABA sales

We observed positive associations between ambient concentrations of atmospheric pollutants and SABA sales for subjects < 40 years old. These are expressed with latency periods of 4 to 10 days and do not tend to increase or decrease according to SES.

This study is the first to examine the relations between exposure to urban air pollution and SABA sales. The use of this indicator, obtained from the four primary French health insurance funds, allowed us to cover > 90% of the local population and to capture the entire range of SES in the SMA. People not covered by these funds are mainly employees of various sectors once publicly owned (railway, electricity, gas), with jobs ranging from manual workers to administrators and mainly in the middle classes. The large number of SABA sales allowed us to measure their associations with air pollution modeled by small areas. This resolution is particularly pertinent for studying this risk factor because its spatial distribution varies strongly within urban areas (Laurent et al., 2008). These large numbers also allowed us to test the existence of interactions by neighborhood SES with satisfactory statistical power. The event analyzed is the patient's purchase of one (or sometimes more) box of drugs, and not a quantity of active ingredient delivered or really inhaled. Naureckas et al., (Naureckas et al., 2005) nonetheless showed that this indicator is a good predictor of the risk of emergency department visits and of hospitalization for asthma attacks in the days immediately afterwards. These purchases generally reflect asthma morbidity less severe than that requiring hospitalization or emergency treatment (Naureckas et al., 2005). A large portion of SABA sales anticipate the respiratory disorders the drugs are intended to treat. These sales, which need not be associated in time with asthma, add some "noise" to the data. Noise is standard in ecologic studies of the

short-term health effects of air pollution because of the influence of unmeasured competing factors (other than air pollution) on the temporal distribution of the health outcomes studied. Nevertheless, if the signal-to-noise ratio is sufficiently high, the effect specifically due to air pollution can be observed. In our study, despite the additional noise due to anticipatory sales, this ratio appears to be high enough to detect statistically significant associations. These associations are consistent with those reported by most panel studies (von Klot et al., 2002; Romeo et al., 2006; Schildcrout et al., 2006) that have investigated the relation between air pollution and SABA consumption.

The associations observed involved latency periods of 4 to 10 days, an order of magnitude similar to the delayed responses observed in two earlier temporal ecologic studies of the relation between air pollution and drug sales mucolytic and antitussive agents (Zeghnoun et al., 1999) "cough and cold preparations," and all types of anti-COPD/antiasthma drugs (Pitard et al., 2004). This is probably because the latency periods between exposure to air pollution and drug purchases result from a mixed process involving both pathophysiological response and management of medicine supplies. The literature (von Klot et al., 2002; Rabinovitch et al., 2006; Romeo et al., 2006; Schildcrout et al., 2006) shows that in people with asthma, air pollution induces respiratory disorders expressed by increased SABA consumption, with low latency periods (several hours 34 to several days (von Klot et al., 2002; Schildcrout et al., 2006)). This increased consumption requires a successive -but not necessarily immediate- replenishment of the SABA supply. Several days of delay, possibly marked by a return to a normal rhythm of use, may pass before the purchase, which may explain the particularly long lags (up to 10 days) observed. For lag 2, we observed odds ratios significantly less than one. Several authors, in other settings and with different methods, report similar findings. Zeghnoun et al. (Zeghnoun et al., 1999) also observed low associations for lag 2, especially for mucolytic and antitussive drugs. Moreover, in a panel study of SABA consumption, Rabinovitch et al. (Rabinovitch et al., 2006) also observed relative risks less than one for the same lag. Von Klot et al., 2002) report comparable observations specifically for lag 1. No satisfactory explanation has yet been found for these observations.

SES did not influence the relation between SABA sales and ambient air pollution. This is consistent with the results appearing above for an indicator of more severe asthma morbidity, interventions of mobile emergency medical services for asthma attacks (Laurent et al., 2008).

In conclusion, emergency calls for asthma attacks SABA sales for children, adolescents and young adults were positively (not significantly for the former, but significantly for SABA sales) associated with PM<sub>10</sub>, NO<sub>2</sub> but not O<sub>3</sub> concentrations modeled by small areas. Small area deprivation did not influence these associations. Nonetheless, discrepancies between our results on emergency calls and those of the study of Lin et al. (Lin et al., 2004) emphasize the need to investigate this question further in other study settings. Similarly, the observations we made regarding SABA sales do not rule out the possibility that SES might be an interaction factor in other settings, for the distribution according to SES of other factors that might modulate the relations between air pollution and asthma morbidity may well differ between countries or even cities. The results on SABA are consistent with those of panels of asthma patients and their SABA consumption, although expressed here with longer time lags. Our results support the usefulness of SABA sales for the analysis of relations between asthma morbidity and air pollution.

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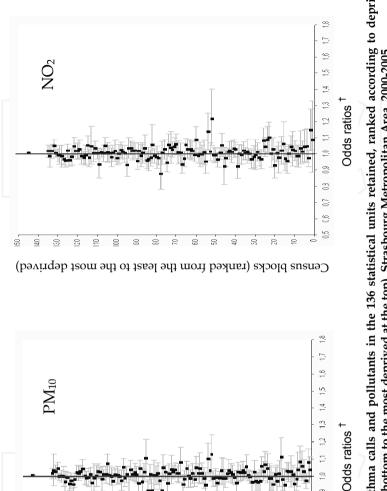


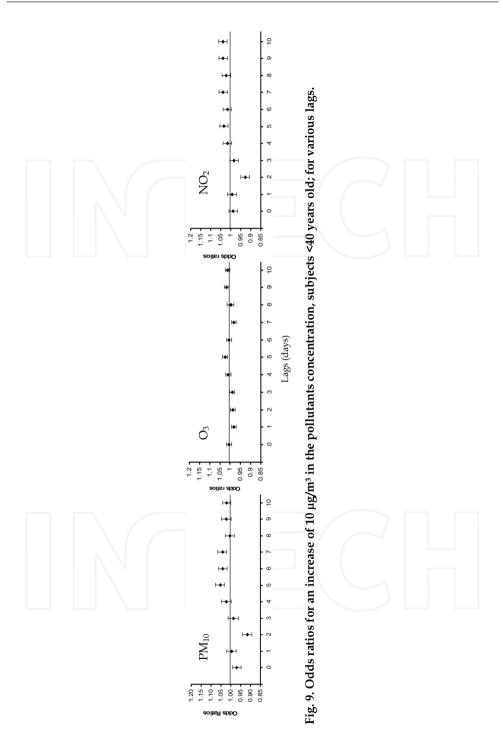
Fig. 8. Odds ratios between asthma calls and pollutants in the 136 statistical units retained, ranked according to deprivation relative humidity, atmospheric pressure, holidays, influenza epidemics and pollen counts. PM10, particulate matter less than 10 Odds ratios reported for a 1 µg.m-3 increase in pollutant concentrations (for the sake of figure visibility), adjusted for temperature, from the least deprived on the bottom to the most deprived at the top). Strasbourg Metropolitan Area, 2000-2005 micrometers in aerodynamic diameter; NO2: nitrogen dioxide

0,0 8,0 70 0,9

€ 64 . 130

2 8

Census blocks (ranked from the least to the most deprived)



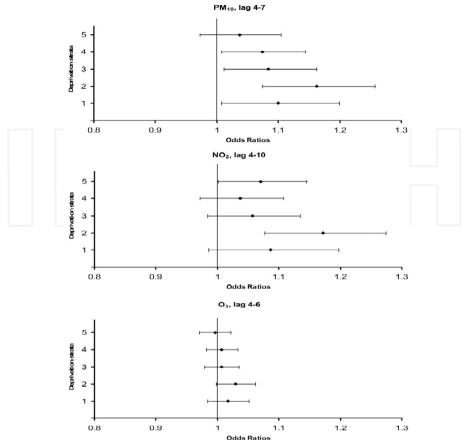


Fig. 10. Interactions by socioeconomic level in those aged < 40 years: associations for the five different strata of socioeconomic levels.

Stratum 1 is the most advantaged and stratum 5 the most deprived. Associations were estimated for "optimal" lags defined according to the associations reported for individual lags, see Figure 7. For example, for  $PM_{10}$  among those younger than 40 years, 4-7 corresponds to the mean concentrations for day D4, D5, D6 and D7.

Table VI. Distribution of SABA sales and air pollutant concentrations, Strasbourg Metropolitan Area, 2004

	Pop	ulation,	age groul	SC	SAB	SABA sales, age gro	age grou	sdi	Pollt	Pollutants (mean	(S,SD)
Deprivation stratum§	Total, 0- 39	6-0	10-19	20-39	Fotal, 0- 39	6-0	10-19	20-39	$ m PM_{10}^{\dagger}$	$NO_2^{\dagger}$	$O_{3^{\ddagger}}$
Stratum 1	43,674	9,342	11,489	22,843	2,140	733	426	981	19.4 9.9	29.9 10.3	63.1 36.5
Stratum 2	47,757	6,359	11,022	27,376	2,538	728	578	1,232	20.5 10.4	34.2 10.2	59.3 36.1
Stratum 3	54,527	9,498	9,550	35,479	2,972	931	460	1,581	21.6 10.6	37.5 10.7	56.0 35.7
Stratum 4	65,994	10,404	11,917	43,673	3,752	1,192	631	1,929	21.4 10.5	37.0 10.6	57.0 35.8
Stratum 5	49,111	12,440	13,269	23,402	3,719	1,332	263	1,624	20.7 10.3	35.7 10.2	58.6 35.8
Overall	261,063	51,043	52,247	152,773	15,121	4,916	2,858	7,347	20.8 10.2	35.0 10.3	58.7 36.0

§ Stratum 1 is the last deprived, and stratum 5 the most deprived

<sup>†</sup>Concentrations averaged for year 2004, in microgram per cubic meter

<sup>‡</sup> Maximum 8-hour daily concentrations, averaged for summer months (April 1 to September 30) of year 2004, in microgram per cubic meter

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Although the climate of the Earth is continually changing from the very beginning, anthropogenic effects, the pollution of the air by combustion and industrial activities make it change so quickly that the adaptation is very difficult for all living organisms. Researcher's role is to make this adaptation easier, to prepare humankind to the new circumstances and challenges, to trace and predict the effects and, if possible, even decrease the harmfulness of these changes. In this book we provide an interdisciplinary collection of new studies and findings on the score of air pollution.

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Phone: +86-21-62489820 Fax: +86-21-62489821