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## Oxolinic Acid Binding at Goethite and Akaganéite Surfaces: Experimental Study and Modeling

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1 **Oxolinic Acid Binding at Goethite and Akaganéite Surfaces: Experimental**  
2 **Study and Modeling**

3  
4  
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24 **Abstract**

25 Oxolinic acid (OA) is a widely used quinolone antibiotic in aquaculture. In this study, its  
26 interactions with synthetic goethite ( $\alpha$ -FeOOH) and akaganéite ( $\beta$ -FeOOH) particle surfaces  
27 were monitored to understand the potential fate of OA in marine sediments where these  
28 phases occur. Batch sorption experiments, liquid chromatography (LC) analyses of  
29 supernatants, Attenuated Total Reflectance-Fourier Transform InfraRed (ATR-FTIR)  
30 spectroscopy and multi-site complexation (MUSIC) modeling were used to monitor OA  
31 binding at these particle surfaces. Both LC and ATR-FTIR showed that adsorption did not  
32 degrade OA, and that OA adsorption was largely unaffected by NaCl concentrations (10-1000  
33 mM). This was explained further by ATR-FTIR suggesting the formation of metal-bonded  
34 complexes at circumneutral to low pHc ( $= -\log [H^+]$ ) and with a strongly hydrogen-bonded  
35 complex at high pHc. The stronger OA binding to akaganéite can be explained both by the  
36 higher isoelectric point/point-of-zero charge (9.6-10) of this mineral than of goethite (9.1-9.4),  
37 and an additional OA surface complexation mechanism at the (010) plane. Geminal sites  
38 ( $\equiv Fe(OH_2)_2^+$ ) at this plane could be especially reactive for metal-bonded complexes, as they  
39 facilitate a mononuclear six-membered chelate complex via the displacement of two  
40 hydroxo/aquo groups at the equatorial plane of a single Fe octahedron. Collectively, these  
41 findings revealed that Fe-oxyhydroxides may strongly contribute to the fate and transport of  
42 OA-type antibacterial agents in marine sediments and waters.

43

## 44 1. Introduction

45 Oxolinic Acid (OA) belongs to a family of quinolone antibiotics that is commonly  
46 used in aquaculture as a prophylactic, or as a chemotherapy agent.<sup>1,2</sup> Unabsorbed or excreted  
47 OA released into waters and/or accumulated in marine and freshwater sediments poses  
48 potential risks to public health and to aquatic life.<sup>3</sup> Due to its frequent use in fish farming,  
49 concentrations of OA can be of the order of 0.01–2.50 ppm in fresh or saline surface waters,  
50 0.01–2.31 ppm in bottom waters, and even 1.81–426.31 ppm in pond sediments.<sup>1,4,5</sup>  
51 Moreover, Greek fish farms release 50 kg of OA annually to the Mediterranean Sea.<sup>4</sup>

52 In marine or freshwater sediments, soils and aquatic environments, the transport and  
53 mobility of quinolone antibiotics are strongly related to reactive oxide minerals through  
54 adsorption and redox reactions.<sup>6-14</sup> In a study on oxolinic acid sorption on marine sediments  
55 Pouliquen and Le Bris<sup>15</sup> concluded that while OA was not transformed in the sediments by  
56 abiotic or biotic processes, it can be sorbed to sediments containing reactive mineral surfaces.  
57 Among these reactive surfaces, iron oxyhydroxides, including goethite ( $\alpha$ -FeOOH) and  
58 akaganéite ( $\beta$ -FeOOH), are ubiquitous in marine environments.<sup>16,17</sup> Akaganéite is a chlorine-  
59 bearing iron oxyhydroxide with a hollandite-type (tunnel-like) structure containing Cl<sup>-</sup> ions  
60 residing in 4 × 4 nm wide tunnels running along the length of acicular nano-sized particles.  
61 This phase can be generally found in environments rich in Fe(II) and Cl<sup>-</sup>, and as a natural  
62 corrosion product in high salinity environmental settings.<sup>18-21</sup> Goethite is the most  
63 thermodynamically stable iron oxyhydroxide mineral and is reported to form through the  
64 transformation of metastable phases such as ferrihydrite, schwertmannite and akaganéite.<sup>22</sup>  
65 Goethite was also shown to be the most common diagenetic iron oxyhydroxide in both marine  
66 and lake sediments.<sup>23</sup>

67 Although OA has been extensively used in industrial fish farming, and thus is  
68 frequently detected in surrounding waters or sediments, little is known about its molecular

69 interactions with suspended particles or mineral constituents in sediments, which is necessary  
70 to accurately assess its fate in aquatic environments. In this study, interactions of OA with  
71 goethite and akaganéite particles were studied by batch experiments, surface complexation  
72 modeling, and vibration spectroscopic investigations. Effects of pHc ( $4 < -\log [H^+] < 11$ ),  
73 salinity (10-1000 mM NaCl) and surface loading on OA binding mechanisms were resolved,  
74 and implications for the fate of OA in marine environments were proposed.

75

## 76 **2. Materials and methods**

### 77 **2.1. Chemicals**

78 All chemicals used were of pro-analytical quality or better and were purchased from  
79 Sigma-Aldrich. An OA (purity > 99%) stock solution was prepared by dissolving 30 mg (115  
80  $\mu$ moles) OA in 20 mL of 1 M NaOH, then diluted to 1 L with ultrapure water.

81

### 82 **2.2. Synthesis and characterization of goethite and akaganéite particles**

83 Goethite and akaganéite particles were synthesized as described in previous  
84 studies.<sup>24,25</sup> The synthesis procedure is given in detail in the supporting information (SI). The  
85 obtained particles were characterized by particle size and B.E.T. specific surface area  $N_2(g)$   
86 adsorption measurements, chemical analysis, transmission electron microscopy (TEM) and X-  
87 ray diffraction (XRD). The B.E.T. specific surface area of the synthetic goethite and  
88 akaganéite were 95 and 239  $m^2/g$ , respectively. Other characteristic properties of these solids  
89 are summarized in the SI.

90

### 91 **2.3. Batch experiments and liquid chromatography analysis**

92 Because high background electrolyte solutions can affect pH measurements,<sup>27</sup> the pH  
93 electrode was calibrated to measure the molarity of the proton ( $-\log [H^+]$ , noted pHc) using

94 solutions of known  $[H^+]$  ( $10^{-5}$ - $10^{-3}$  M) at different  $[NaCl]$ , and the results are provided in this  
95 form in this study.

96 OA solubility experiments (undersaturation direction) were conducted by suspending  
97 solid OA (~3-5 mg) in 10 mL of solution in 100 mM NaCl as a function of pHc. The  
98 suspensions were equilibrated for 24 h, after which the supernatants were filtered (0.2  $\mu$ m)  
99 and OA concentrations were measured with high pressure liquid-chromatography coupled  
100 with UV detection (HPLC-UV). As discussed later, OA solubility is relatively low (~10  $\mu$ M)  
101 at acidic pHc, in agreement with previous studies.<sup>26</sup>

102 An akaganéite (stock) suspension was prepared as previously reported.<sup>25</sup> Aliquots of  
103 the original stock suspension (21 g/L, 5023 m<sup>2</sup>/L) were used to prepare four suspensions (2.1  
104 g/L, 502.3 m<sup>2</sup>/L) in different concentrations of NaCl (10, 100, 450 or 1000 mM). These  
105 suspensions were equilibrated for over 30 days to ensure complete ion exchange between bulk  
106 akaganéite and solution ions, as previously explained by Kozin and Boily.<sup>25</sup> Goethite  
107 suspensions were prepared at the chosen ionic strength just before adding OA.

108 Batch adsorption experiments were performed under conditions of varied reaction  
109 time, OA concentration, pHc and ionic strength. Kinetic studies were conducted for pHc = 6,  
110  $[NaCl] = 100$  mM,  $[OA]_{tot} = 50$   $\mu$ M with an equivalent surface area of solids, i.e. 50 m<sup>2</sup>/L of  
111 goethite or akaganéite. All batch experiments were performed under an atmosphere of N<sub>2</sub>(g)  
112 to purge dissolved CO<sub>2</sub> from the aqueous solutions. Briefly, solutions of 1 to 100  $\mu$ M OA  
113 were mixed with goethite or akaganéite (50 m<sup>2</sup>/L) at the desired concentration of NaCl (10,  
114 100, 450 or 1000 mM). pHc was thereafter adjusted to the desired value by addition of 0.1 M  
115 HCl solutions (the initial pHc value was about 10). In another set of experiments, adsorption  
116 isotherms were collected at pHc = 6.1 $\pm$ 0.1, for  $[OA]_{tot}$  concentrations in the 1-50  $\mu$ M range  
117 for goethite, and the 3-100  $\mu$ M range for akaganéite. All suspensions were then equilibrated  
118 for 24 h and the supernatants filtered (0.2  $\mu$ m) prior to soluble OA determination by HPLC-

119 UV. Desorption tests (pHc=11) were conducted under various experimental conditions as a  
120 means to verify the mass balance of OA in the systems under study.

121 Soluble OA concentrations were determined by a Waters 600 controller high  
122 performance liquid chromatography (HPLC) system equipped with an auto sampler (Waters  
123 717 plus), a C18 column (250 mm×4.6 mm i.d., 5 μm) and a UV detector (275 nm, Waters  
124 2489). The mobile phase was a mixture of water/acetonitrile (60:40 v/v) containing 0.1% of  
125 formic acid. The flow rate of the mobile phase was set at 1 mL/min in the isocratic mode.  
126 Finally, the complete experimental solubility and sorption datasets are provided in SI (Tables  
127 S1, S2 and S3).

128

#### 129 **2.4. ATR-FTIR spectroscopy**

130 Attenuated total reflectance-Fourier transform infrared (ATR-FTIR) spectra were  
131 recorded between in the 780-4000 cm<sup>-1</sup> region on an IS50 Nicolet spectrometer equipped with  
132 a KBr beamsplitter and a liquid Nitrogen cooled MCT (Mercury Cadmium Telluride)  
133 detector. A nine-reflection diamond ATR accessory (DurasamplIR™, SensIR Technologies)  
134 was used for acquiring spectra of wet samples. The resolution of the single beam spectra was  
135 4 cm<sup>-1</sup>.

136 Three series of experiments were conducted with akaganéite for: (i) 5 < pH < with  
137 [OA]<sub>tot</sub>=50 μM and [NaCl]=100 mM, (ii) 10 ≤ [NaCl] ≤ 1000 mM with [OA]<sub>tot</sub>=50 μM and  
138 pHc=6 and (iii) 20 ≤ [OA]<sub>tot</sub> ≤ 50 μM with pHc=6 and [NaCl]=100 mM. For goethite, the  
139 same experiments were conducted except for pHc < 7, where [OA]<sub>tot</sub>=20 μM because OA  
140 partially precipitated at 50 μM (see below). Prior to ATR-FTIR analysis, tubes from batch  
141 sorption experiments were centrifuged at 10,000 g for 30 min. Wet mineral pastes were  
142 directly and uniformly applied to the diamond ATR crystal then covered with a lid of a flow-  
143 through cell to prevent the evaporation of water. ATR-FTIR spectra were then recorded



144 immediately. Reference spectra include those for pure water, the filtered (0.2  $\mu\text{m}$ ) supernatant  
145 from each batch adsorption experiment, as well as  $\text{OA}_{(s)}$  in its solid form but covered with a  
146 drop of water to apply it more uniformly. A reference spectrum of aqueous OA was also  
147 acquired from a 10 mM OA (2.5 g/L) solution in 1 M NaOH. This solution ensured a  
148 sufficiently high concentration of  $\text{OA}^-_{(aq)}$  for ATR-FTIR analysis. Otherwise, OA was not  
149 detected in any of the supernatants given its low solubility. All spectra are shown for the  
150 fingerprint 1200–1800  $\text{cm}^{-1}$  region only. Additional ATR-FTIR measurements of OA in  $\text{D}_2\text{O}$   
151  $\text{OA}^-_{(aq)}$ , OA-goethite (pD = 7.81) and OA-akanag ite (pD = 8.51) in  $\text{D}_2\text{O}$  were also carried  
152 out to resolve bands arising from moieties that are directly affected by exchange or hydrogen-  
153 bonding with water. Sample preparation is detailed further in the Supporting Information (SI).  
154

## 155 2.5. Surface complexation modeling

156 Predictions of sorption vs pHc were made using the multi-site complexation (MUSIC)  
157 model approach.<sup>28</sup> The geochemical speciation code PHREEQC (version 2)<sup>29</sup> was used for  
158 surface complexation calculations. PHREEQC includes the three plane model (TPM).  
159 Charges of the adsorbates were distributed among the 0 ( $\text{H}^+$ , metal-bonded OA), 1 (hydrogen-  
160 bonded OA), and 2 ( $\text{Na}^+$ ,  $\text{Cl}^-$ ) planes of the TPM. The TPM requires two capacitances for the  
161 0- and the 1-plane,  $C_1$  and  $C_2$ , respectively (see SI for more details). PhreePlot<sup>30</sup> was used to  
162 estimate parameters for OA sorption to goethite and akagan ite. Using a modified Marquardt-  
163 Levenberg procedure,<sup>31</sup> PhreePlot provides a statistical uncertainty for the estimated  
164 parameters, which are reported in this study. For goethite or akagan ite, different types of  
165 surface sites can exist depending on the proportion of crystal planes, which should have  
166 different binding affinities. All information concerning the dominant crystal planes of  
167 particles and corresponding site densities are given in the SI. Charge densities and  
168 corresponding electrostatic interactions at particle termination (i.e. (021) plane of goethite and

169 (010) plane of akaganéite) were computed individually, as well as formation constants of all  
170 surface species are reported in Table 1. Surface complexation constants reported for OA were  
171 corrected according to Venema *et al.*<sup>32</sup> (see also<sup>33</sup>).

172

### 173 **3. Results and Discussion**

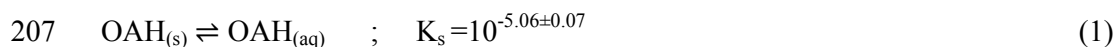
#### 174 **3.1. Batch experiments**

175 Kinetic batch adsorption experiments conducted over a 7-day period showed that  
176 equilibrium was achieved within 2 h of reaction time (Figure S1). Desorption tests also  
177 showed that OA was removed only by adsorption or precipitation, and that breakdown by, for  
178 example, oxidation, did not occur under the experimental conditions of this study.

179 Adsorption curves for 10  $\mu\text{M}$  OA solutions reacted in 50  $\text{m}^2/\text{L}$  mineral suspensions  
180 (Fig. 1a,b) revealed greater OA loadings on akaganéite than on goethite. At  $\text{pH}_c < 6.5$ , only  
181  $\sim 70\%$  of OA sorbs onto goethite (Fig. 1a, Table S1) whereas an almost complete uptake was  
182 achieved on akaganéite at  $\text{pH}_c=7.5$  ( $\sim 98\%$ ; Fig. 1b, Table S2). OA adsorption was greatest  
183 under acidic to circumneutral pH, and lowest under alkaline conditions, as typically  
184 encountered for anionic ligands. Furthermore, an insignificant effect of NaCl concentration  
185 was detected, so OA binding should predominantly occur as strong metal-bonded complexes  
186 with surface Fe sites and/or directly hydrogen-bonded complexes with surface hydroxo  
187 groups. Resistance of the latter to variations in ionic strength was notably demonstrated by  
188 Johnson *et al.*<sup>34</sup>. Weaker outer-sphere complexes should, on the other hand, be more affected  
189 by variations in ionic strength.<sup>35</sup> This observation is consistent with the adsorption of  
190 flumequine (analogous structure of OA) to goethite where the effect of ionic strength was also  
191 negligible ( $1 < [\text{NaCl}] < 100 \text{ mM}$ ).<sup>14</sup> Our explanation for the greater OA loadings achieved in  
192 akaganéite is two-fold. First, we note that while reactive site density of these two minerals  
193 under study are highly comparable (see details in SI), the point-of-zero charge of akaganéite

194 (9.6-10)<sup>25</sup> is higher than that of goethite (9.1-9.4)<sup>24,36</sup>. Second, geminal  $\equiv\text{Fe}(\text{OH})_2^+$  groups of  
195 the (010) plane of akaganéite are expected to be strongly reactive sites for binding anionic  
196 species.<sup>25,37,38</sup> The site density of these groups is estimated at 0.18 site/nm<sup>2</sup> (i.e. 5%  $\times$  3.53  
197 site/nm<sup>2</sup>), which may favor an additional OA removal mechanism, as will be suggested  
198 further by modeling (Fig. 1b and 1d).

199 In case of goethite (pHc < 7), the plateau is found higher for  $[\text{OA}]_{\text{tot}}=50 \mu\text{M}$  (~85%)  
200 than for  $[\text{OA}]_{\text{tot}}=10 \mu\text{M}$  (~70%) (Fig. 1a and 1c, Table S1), i.e. OA loading is more than 5  
201 times larger at 50  $\mu\text{M}$  than at 10  $\mu\text{M}$ , which may suggest an additional removal mechanism  
202 such as surface or bulk precipitation of OA in this pHc range. The latter possibility was tested  
203 by conducting solubility experiments of OA at different pHc values (Fig. 2a, Table S3).  
204 Additionally, the OA solubility was calculated with PHREEQC using the acidity constant  
205 ( $K_a$ ) value from the IUPAC stability constant database<sup>39</sup> and the solubility constant ( $K_s$ ) fitted  
206 to our data with Phreeplot:



209 The experimental solubility values are in good agreement with both calculated (Fig. 2a) and  
210 literature data (e.g. 23  $\mu\text{M}$  at pH = 7).<sup>26</sup> The equilibrium OA concentrations ( $[\text{OA}]_{\text{aq}}$ ) in the  
211 presence of goethite or akaganéite at  $[\text{OA}]_{\text{tot}}=50 \mu\text{M}$  are also plotted in Figure 2a. In contact  
212 with akaganéite,  $[\text{OA}]_{\text{aq}}$  is below the solubility limit for all pHc tested, ruling out the bulk  
213 precipitation in this experimental set.  $[\text{OA}]_{\text{aq}}$  values in goethite are however closer to the  
214 solubility curve at pHc < 7.5, a result possibly explaining the larger loadings achieved at  
215  $[\text{OA}]_{\text{tot}}=50 \mu\text{M}$ . Accordingly,  $[\text{OA}]_{\text{aq}}$  values measured at low ligand loadings ( $[\text{OA}]_{\text{tot}}=10$   
216  $\mu\text{M}$ ) are below the OA solubility limit (Fig. 2a).

217 Adsorption isotherms (Fig. 2b) add further validation to solubility data where a sharp  
218 rise in OA removal with goethite occurred at pHc=6.1 $\pm$ 0.1 and  $[\text{OA}]_{\text{aq}} \approx 10 \mu\text{M}$ ,

219 corresponding to the solubility limit of OA. The OA adsorption isotherm for akaganéite does  
220 not, on the other hand, reach this solubility limit, but rather reaches a maximal loading  
221 corresponding to  $Q_{\max}=1.14$  OA/nm<sup>2</sup>. This value was obtained by modeling OA-akaganéite  
222 data with the Langmuir adsorption isotherm:

223

$$224 \quad [OA]_{\text{sorbed}} = Q_{\max} \frac{K_L \times [OA]_{\text{aq}}}{1 + K_L \times [OA]_{\text{aq}}} \quad (3)$$

225

226 with an adsorption constant of  $K_L = 0.53 \mu\text{M}^{-1}$  (see the fit with the Langmuir equation in Fig.  
227 S2). As this  $Q_{\max}$  value is about one-third our estimated reactive site density at FeOOH  
228 surfaces (*i.e.*  $\sim 3$  -OH/nm<sup>2</sup>), OA surface complexes binding may realistically involve two or  
229 three -OH sites. The molecular-level nature of these interactions will however be addressed in  
230 greater detail in the following section.

231

## 232 3.2. ATR-FTIR Spectroscopic investigations

### 233 3.2.1. $OA_{(s)}$ and $OA_{(aq)}^-$ in 1 M NaOH

234 The ATR-FTIR spectra of solid and soluble OA (Fig. 3) were used as a basis for  
235 assigning surface-bound OA species. Owing to the low solubility of OA, band assignments in  
236 the soluble species were made from a 1 M NaOH solution (Fig. 3).

237 The absorption band of  $OA_{(s)}$  and  $OA_{(aq)}$  at 1040 cm<sup>-1</sup> (see Figure S3) corresponds to  
238 the stretching modes of the ether group ( $\nu_{\text{C-O-C}}$ ). The FTIR spectra of  $OA_{(s)}$  and  $OA_{(aq)}$  exhibit  
239 three bands with maximum absorbance between 1470 and 1510 cm<sup>-1</sup>. These bands are  
240 assigned to C=C stretches ( $\nu_{\text{C=C,ring}}$ ) and C-H bends ( $\delta_{\text{C-H,ring}}$ ) in the aromatic and quinolinone  
241 rings.<sup>40</sup> The intense band of  $OA_{(s)}$  at 1440 cm<sup>-1</sup>, which disappears in the spectrum of  $OA_{(aq)}$ ,  
242 may be attributed to perturbation of  $\nu_{\text{C=C,ring}}$  or  $\delta_{\text{C-H,ring}}$  due to intermolecular interaction in  
243  $OA_{(s)}$ . The major difference between ATR-FTIR spectra of  $OA_{(s)}$  and  $OA_{(aq)}$  lies in the

244 disappearance of two bands at 1705 and 1302  $\text{cm}^{-1}$  from the deprotonation of the carboxyl  
245 groups, *i.e.* disappearance of C=O and C-OH stretching modes of the carboxylic group. This  
246 deprotonation gives, in turn, rise to the 1395 and 1586  $\text{cm}^{-1}$  bands, which are respectively  
247 assigned to symmetric ( $\nu_{\text{COO},s}$ ) and asymmetric ( $\nu_{\text{COO},as}$ ) stretching modes of the carboxylate,  
248 and where  $\Delta\nu = \nu_{\text{COO},as} - \nu_{\text{COO},s} = 191 \text{ cm}^{-1}$ . These values are, moreover, consistent with  
249 previous studies on (fluoro)quinolones, ciprofloxacin (CIP) assigning  $\nu_{\text{COO},s}$  to a band at  
250 1360-1380  $\text{cm}^{-1}$ ,<sup>6-11</sup> and on ofloxacin (OFX) with a similar band at 1340  $\text{cm}^{-1}$ .<sup>12-14</sup> Our  
251 spectrum for  $\text{OA}_{(aq)}$  also showed a distinct band at 1340  $\text{cm}^{-1}$ , but for reasons given later, it is  
252 not assigned to  $\nu_{\text{COO},s}$ . The absorption band at 1635  $\text{cm}^{-1}$  is assigned to the stretching of the  
253 quinolinone ring carbonyl group ( $\nu_{\text{C=O,carbonyl}}$ ), in accordance with previous spectroscopic  
254 studies dedicated to (fluoro)quinolones.<sup>6-14,41</sup>

255

### 256 3.2.2. OA-goethite

257 Figure 3a shows the effect of pHc on the ATR-FTIR spectra of 20  $\mu\text{M}$  OA (for pHc <  
258 7) or 50  $\mu\text{M}$  OA (for pHc >7) reacted with 50  $\text{m}^2/\text{L}$  goethite in 100 mM NaCl. As shown in  
259 Figure S3a, the characteristic features of both  $\text{OA}_{(s)}$  and  $\text{OA}_{(aq)}$  were observed for OA-goethite  
260 samples at pHc = 6.12 for  $[\text{OA}]_{\text{tot}}=50\mu\text{M}$ . Only the characteristic of  $\text{OA}_{(aq)}$  appear for pHc  $\approx$  6  
261 with  $[\text{OA}]_{\text{tot}} \leq 30 \mu\text{M}$  and for pH > 7 with  $[\text{OA}]_{\text{tot}} = 50 \mu\text{M}$  (Fig. 2, Fig. S3), in agreement  
262 with our batch results. Especially, the absence of band at 1705  $\text{cm}^{-1}$  suggests the presence of  
263 deprotonated carboxylate groups.

264 At pHc = 5.20 and 6.39 ( $[\text{OA}]_{\text{tot}}=20\mu\text{M}$ ), the band at 1395  $\text{cm}^{-1}$  for  $\text{OA}_{(aq)}$  almost  
265 completely disappears and is replaced by a band centered at 1340  $\text{cm}^{-1}$ . This further supports  
266 the assignment of  $\nu_{\text{COO},s}$  to the band at 1395  $\text{cm}^{-1}$  in  $\text{OA}_{(aq)}$ . The  $\nu_{\text{COO},as}$  band might be slightly  
267 blue-shifted (1590  $\text{cm}^{-1}$ ), which gives  $\Delta\nu \approx 250 \text{ cm}^{-1}$ , a value suggesting that only one oxygen  
268 of the carboxylate binds to a surface Fe.<sup>42</sup> A blue shift of  $\nu_{\text{C=O,carbonyl}}$  (1642  $\text{cm}^{-1}$ ) compared to

269  $\text{OA}_{(\text{aq})}$  or  $\text{OA}_{(\text{s})}$ , also suggests the involvement of the keto group in the surface complex. A  
270 similar shift from protonated  $\text{OA}_{(\text{s})}$  to  $\text{Fe}(\text{OA})_{3(\text{s})}$  was previously reported, where OA forms a  
271 bidentate deprotonated complex with  $\text{Fe}^{3+}$  through the keto group and one carboxylate  
272 oxygen.<sup>41</sup> The effects of NaCl concentration were also investigated for 20  $\mu\text{M}$  OA and 50  
273  $\text{m}^2/\text{L}$  goethite at  $\text{pHc} \approx 6$  (Fig. S3b). The weak ionic strength dependence on OA loadings and  
274 ATR-FTIR spectra point to the predominance of OA species that are inner-spherically bound  
275 to surface Fe sites (*i.e.* metal-bonded complex; MB), involving one oxygen of the carboxylate  
276 and the oxygen of the keto group of OA. The involvement of both the carboxylate and keto  
277 groups was proposed in previous studies dedicated to other quinolones binding to hydrous  
278 ferric oxides<sup>6</sup>, magnetite<sup>9</sup>, hematite<sup>11</sup> and goethite<sup>10,14</sup> or observed for metal cation-OA  
279 complexes in water.<sup>41</sup> The involvement of both the carboxylate and the phenolate of salicylate  
280 and hydroxynaphthoate has also been evidenced at goethite surfaces.<sup>43-46</sup> We cannot directly  
281 conclude from the ATR-FTIR spectra whether this complex involves one (Fig. 4a) or two  
282 surface Fe atoms (Fig. 4b). Ideally, the steric constraints at the (101)/(001) as well as the  
283 (210) planes (Pnma group) should promote bridging between two Fe atoms separated by 3 Å  
284 from one another (Fig. 4b). Moreover, defects ought to expose additional possibilities for  
285 mononuclear bidentate binding (*cf.* six-membered chelate complexation, Fig. 4a), namely at  
286 equatorial edges of a Fe octahedron.<sup>47</sup>

287 At  $\text{pHc} = 7.60$  and  $8.13$  ( $[\text{OA}]_{\text{tot}}=50\mu\text{M}$ ), all bands remain unaffected compared to  
288  $\text{pHc} = 5.20$  and  $6.39$  ( $[\text{OA}]_{\text{tot}}=20\mu\text{M}$ ), except the band at  $1340\text{ cm}^{-1}$ , which broadens and  
289 whose full width at half maximum considerably increases with  $\text{pHc}$ . This band broadening is  
290 taken as an indication of interactions of carboxylic groups with surface hydroxo groups with  
291 various hydrogen-bond strengths (*cf.* inhomogeneous band broadening<sup>48</sup>). An ATR-FTIR  
292 spectrum of OA-goethite recorded in  $\text{D}_2\text{O}$  for  $\text{pD} = 7.81$  ( $[\text{OA}]_{\text{tot}}=50\mu\text{M}$ ) reveals that the  
293  $1340\text{ cm}^{-1}$  band is the only one that is red-shifted in relation to the counterpart spectrum in

294 H<sub>2</sub>O (Fig. S4), thus suggesting close interactions between carboxyl groups and O-D/H  
295 moieties. These observations could thus point to a hydrogen-bonded (HB) complex involving  
296 a direct hydrogen bond between carboxyl and keto moieties and surface (hydr)oxo groups.  
297 This ionic strength independent complex thus contrasts with outer-sphere complexes  
298 involving hydration waters.<sup>34</sup> The increased intensity of the band at 1340 cm<sup>-1</sup> suggests that a  
299 strong form of hydrogen bonding is favored at high pHc whereas metal-bonded complexation  
300 is favored at low pHc. This interpretation is in line with, for example, phthalate or  
301 hydroxynaphthoate sorption to goethite<sup>44,46,47</sup> where HB complexes are involved with two –  
302 OH sites at the (101)/(001) planes (Fig. 4c).

303

### 304 3.2.3. OA-akaganéite

305 Figure 3b shows the effect of pHc on ATR-FTIR spectra of 50 μM OA interacting  
306 with 50 m<sup>2</sup>/L akaganéite in 100 mM NaCl. In agreement with our batch sorption results (Fig.  
307 2), the characteristic features of OA<sub>(s)</sub> are absent and therefore spectra must be attributed to  
308 the surface complexed OA only. This is further confirmed by the analysis of OA-akaganéite  
309 samples at lower [OA]<sub>tot</sub> (down to 20 μM) at pHc ≈ 6 in 100 mM NaCl (Fig. S3c). The  
310 spectra recorded at pHc = 5.24 and 6.20 are similar to those of OA-goethite when the amount  
311 of precipitated OA is negligible, namely  $\nu_{\text{C=O,carbonyl}} = 1642 \text{ cm}^{-1}$  and  $\nu_{\text{COO}_s} = 1330 \text{ cm}^{-1}$ . The  
312 intense band observed for OA<sub>(aq)</sub> at 1395 cm<sup>-1</sup> disappears in the spectra recorded for OA-  
313 akaganéite. The band of  $\nu_{\text{COO,as}}$  (1590 cm<sup>-1</sup>) is also slightly blue-shifted compared to OA<sub>(aq)</sub>.  
314 As for goethite, no significant effect of NaCl concentration on OA sorption to akaganéite can  
315 be observed for pHc ≈ 6 (Fig. 1b), in accordance with the ATR-FTIR spectra (Fig. S3d).  
316 Because these experimental conditions correspond to relatively high OA loadings (~0.5  
317 OA/nm<sup>2</sup>), formation of a bridging complex between two surface Fe involving one oxygen of  
318 the carboxylate and the oxygen of the keto group of OA (Fig. 4b) is likely expected at the

319 (100)/(001) plane. In addition, the effect of OA surface coverage is investigated for  $[OA]_{tot}$   
320 down to 20  $\mu\text{M}$  (i.e. 0.24 adsorbed OA molecule per  $\text{nm}^2$ ) at  $\text{pHc} \approx 6$  and 100 mM NaCl (Fig.  
321 S3c). At such loading, OA surface complexation to the (010) plane of akaganéite is expected  
322 to be the major sorption mechanism (Fig. 2b, Fig. S2). No change in the coordination mode of  
323 OA on akaganéite can be derived from our ATR-FTIR spectra when  $[OA]_{tot}$  varies. We  
324 propose the formation of a MB surface complex involving one oxygen of the carboxylate and  
325 the oxygen of the keto group of OA. Because both water molecules of the  $\equiv\text{Fe}(\text{OH}_2)_2^+$  groups  
326 can be released during the complexation of OA, we propose the formation of a six-membered  
327 chelate, involving only one Fe (Fig. 4a). Therefore, the proposed structure for the OA  
328 complex at the akaganéite (010) is presently based on crystallographic considerations.

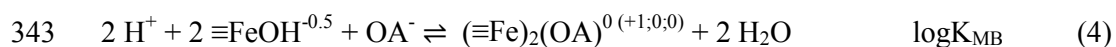
329 At high pHc, we propose the predominance of a HB complex seen through a band at  
330  $\sim 1350 \text{ cm}^{-1}$ . This band is broad and overlaps with the band at  $1330 \text{ cm}^{-1}$ , especially for pHc =  
331 8.02 and 8.66. Its disappearance in  $\text{D}_2\text{O}$  (Fig. S4) suggests that it is sensitive to the hydrogen  
332 bonding environment of OA, as in the case of goethite. As no significant ionic strength effects  
333 were observed, OA could be directly hydrogen bonded to surface (hydr)oxo groups of  
334 akaganéite (Fig. 4c, Fig. S3d).

335

### 336 3.3. Surface complexation modeling

337 According to the aforementioned crystallographic considerations and our  
338 spectroscopic investigations, MB surface complexes at the goethite (101)/(001) and (210)  
339 planes (Pnma group) are expected to be comparable to those on the (100) and (001) planes of  
340 akaganéite. OA forms a complex involving keto group and one oxygen of the carboxylate  
341 group, and it can be expressed in the following reaction:

342



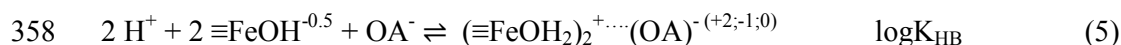


344

345 where (+1;0;0) denotes the charge variations at the 0-, 1- and 2-planes ( $\Delta z_0; \Delta z_1; \Delta z_2$ ). Only  
 346 singly coordinated ( $\equiv\text{FeOH}^{-0.5}$ ) surface sites are considered to be involved in the latter  
 347 reaction given the propensity for ligand exchange of these sites. In the case of goethite, note  
 348 that  $\equiv\text{FeOH}^{-0.5}$  is used as a building block for different adsorption sites on the (101)/(001) and  
 349 the (210) planes. Therefore, Eq. 4 does not denote the number of Fe(III) octahedra in a  
 350 complex, but only the number of  $\equiv\text{FeOH}^{-0.5}$  sites that may or may not be of the same Fe(III)  
 351 octahedron.<sup>47</sup> Our results cannot resolve the number of Fe(III) bound to OA explicitly with  
 352 the data on hand. In two recent MUSIC modeling studies, CIP<sup>10</sup> and OFX<sup>14</sup> were assumed to  
 353 bind only to the goethite (210) plane. However, from our data, we cannot exclude that OA  
 354 binds to the goethite (101)/(001) planes.

355 Hydrogen-bonded (HB) complexation could form by involving various types of  
 356 surface sites. For the sake of simplicity, only singly  $\equiv\text{FeOH}^{-0.5}$  sites are considered:<sup>47</sup>

357



359

360 Preliminary tests showed that  $\log K_{\text{HB}}$  and  $\log K_{\text{MB}}$  values did not significantly differ for  
 361 goethite. Therefore, they were constrained at the same value during the fit. The same strategy  
 362 was applied for akaganéite. According to preliminary tests, it appeared unnecessary to include  
 363 an additional charge distribution (CD)<sup>28</sup> term, unlike previous studies dedicated to CIP<sup>10</sup> and  
 364 OFX<sup>14</sup> that can bind to mineral surfaces under the zwitterion form.

365 Finally, to account for OA binding on the Fe sites of the (010) akaganéite plane, an  
 366 additional reaction was used, which we here describe as a geminal hydroxo/aquo  
 367 displacement reaction:

368



370

371 In all of our modeling attempts, we neglect  $Cl^-$  surface complexation to the akaganéite (010)  
372 plane highlighted in earlier work<sup>25,37,38</sup> as our data reveal no ionic strength dependence on OA  
373 sorption.  $Cl^-$  surface complexation to the akaganéite (010) plane may be too weak to compete  
374 significantly with OA under the present conditions.

375 Surface complexation reactions and parameters for OA are reported in Table 1.  
376 Results of the fit with PhreePlot are shown in Figure 1 for pH-edges, and in Figure 2b for  
377 sorption isotherms. Our model almost predicts  $[NaCl]$  independent OA loading (Fig. 1a,b).  
378 Loading effects and additional apparent OA uptake on goethite (Fig. 1c), actually due to the  
379 bulk precipitation of OA are also well-predicted. Speciation calculations show that about 50%  
380 of 50  $\mu M$  of OA precipitate in contact with goethite at  $pH_c < 6.5$ , in agreement with  
381 spectroscopic observations for samples prepared with high  $[OA]_{tot}$  and  $pH_c < 7$  (Fig. S3a). In  
382 addition, the fraction of hydrogen-bonded OA at the goethite (101)/(001) planes or akaganéite  
383 (100)/(001) planes increases with  $pH_c$  (Fig. S5), consistently with our FTIR data.  
384 Discrepancies between experimental and modeling results are observed for akaganéite at high  
385 loadings ( $[OA]_{tot} > 20 \mu M$ ; Fig. 1d, Fig. 2a), probably due to steric hindrance and/or  
386 electrostatic effects at high ligand loadings. Nevertheless, the model accurately predicts OA  
387 sorption to akaganéite at environmentally relevant concentrations, i.e. lower OA loadings  
388 ( $[OA]_{tot} < 20 \mu M$ ; Fig. 1b, Fig. 2a).

389 The formation of a strong complex at the akaganéite (010) plane significantly  
390 contributes to the overall uptake, even when the surface is almost saturated with OA. Geminal  
391 sites ( $\equiv Fe(OH_2)_2^+$ ) at this plane could be especially reactive for metal-bonded binding, as they  
392 facilitate a mononuclear six-membered chelate complex via the displacement of two  
393 hydroxo/aquo groups at the equatorial plane of a single Fe octahedron. This form of

394 interaction is generally excluded on the other dominant, yet idealized, (001)/(101)/(210)  
395 planes of goethite and (100)/(001) of akaganéite where bridging metal-bonded and hydrogen-  
396 bonded complexes are expected to be favored. The contribution of this complex is illustrated  
397 e.g. in Figure 1d for  $[OA]_{tot}=50 \mu M$ , where  $\equiv Fe(OA)^0$  represents about 20% of OA uptake.  
398 The model predicts that the latter complex prevails at low loadings (Fig. 2b) and high pHc. In  
399 our experimental conditions, for  $[OA]_{tot}=10 \mu M$ ,  $[NaCl]=100 \text{ mM}$  and  $pHc=8$ ,  $\equiv Fe(OA)^0$   
400 accounts for ~65% of OA uptake on akaganéite (Fig. 1b).

401

#### 402 **4. Environmental implications**

403 Our combined kinetic, liquid chromatographic and ATR-FTIR investigations showed that  
404 only adsorption, and not oxidation, controls OA removal from aqueous suspensions of  
405 goethite and akaganéite. We demonstrated that OA binds to goethite or akaganéite surfaces  
406 through MB bidentate complex formation at circumneutral to low pH and through a strong  
407 HB complex at high pH. The dominant MB complex involved direct binding between two  
408 surface Fe with keto and carboxylate oxygens of OA, predominantly in a bridging-type  
409 complex. This complexation mode was confirmed across a wide range of salinity (10 to 1000  
410 mM NaCl) including seawater-like one (i.e. 599 mM). OA sorption to akaganéite is greater  
411 than to goethite, a result that can be attributed to the higher isoelectric point of this mineral,  
412 and also the formation of a strong complex at the akaganéite (010) plane. The contribution of  
413 this strong binding with geminal sites ( $\equiv Fe(OH_2)_2^+$ ) of the akaganéite (010) plane in the  
414 overall sorption extent are strongly dependent on OA concentration. This sorption behavior  
415 may explain the strong retention of OA observed in sediments containing iron mineral  
416 surfaces,<sup>15</sup> but also to mobile FeOOH particles or colloids present in freshwater and marine  
417 environments. This may also explain the slow transformation under dark conditions or high  
418 persistence of OA in pond waters<sup>49</sup>. These results may therefore have strong implications for

419 the understanding of both colloidal and aqueous transport of antibacterial agents used in  
420 aquaculture.

421

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425

#### 426 **Supporting Information Available**

427 Details of the synthesis and characterization of goethite and akaganéite particles, surface  
428 complexation modeling; additional results for kinetic sorption/desorption, OA sorption  
429 isotherms and ATR-FTIR spectroscopy in H<sub>2</sub>O or D<sub>2</sub>O; tables with the whole experimental  
430 datasets. This information is available free of charge via the Internet at <http://pubs.acs.org/>.

431

#### 432 **References**

433

- 434 1. Le, T. X.; Muneke, Y. Residues of selected antibiotics in water and mud from shrimp  
435 ponds in mangrove areas in Vietnam. *Mar. Pollut. Bull.*, **2004**, 49, 922–929.
- 436 2. Samuelsen, O.B. Pharmacokinetics of quinolones in fish: a review. *Aquaculture*, **2006**,  
437 255, 55–75.
- 438 3. Samuelsen, O.B.; Ervik, A.; Pursell, L.; Smith, P. Single-dose pharmacokinetic study of  
439 oxolinic acid and vetoquinol, an oxolinic acid ester, in Atlantic salmon (*Salmo salar*)  
440 held in seawater and in vitro antibacterial activity against *Aeromonas salmonicida*.  
441 *Aquaculture*, **2000**, 187, 213–224.
- 442 4. Rigos, G.; Nengas, I.; Alexis, M.; Troisi, G.M. Potential drug (oxytetracycline and  
443 oxolinic acid) pollution from Mediterranean sparid fish farms. *Aquat. Toxicol.*, **2004**,  
444 69, 281–288.
- 445 5. Tamtam, F.; Mercier, F.; Le Bot, B.; Eurin, J.; Tuc Dinh, Q.; Clement, M.; Chevreuil,  
446 M. Occurrence and fate of antibiotics in the Seine River in various hydrological  
447 conditions. *Sci. Total Environ.*, **2008**, 393, 84–95.
- 448 6. Gu, C.; Karthikeyan, K.G. Sorption of the Antimicrobial Ciprofloxacin To Aluminum  
449 and Iron Hydrated Oxides. *Environ. Sci. Technol.*, **2005**, 39 (23), 9166–9173.
- 450 7. Trivedi, P.; Vasudevan, D. Spectroscopic Investigation of Ciprofloxacin Speciation at  
451 the Goethite–Water Interface. *Environ. Sci. Technol.*, **2007**, 41 (9), 3153–3158.

- 452 8. Pei, Z.-G.; Shan, X.-Q.; Kong, J.-J.; Wen, B.; Owens, G. Coadsorption of ciprofloxacin  
453 and Cu(II) on montmorillonite and kaolinite as affected by solution pH, *Environ. Sci.*  
454 *Technol.*, **2010**, 44 (3), 915-920.
- 455 9. Rakshit, S.; Sarkar, D.; Elzinga, E.J.; Punamiya, P.; Datta, R. Mechanisms of  
456 ciprofloxacin removal by nano-sized magnetite. *J. Hazard. Mater.*, **2013**, 246-247,  
457 221-226.
- 458 10. Gu, X.; Tan, Y.; Tong, F.; Gu, C. Surface complexation modeling of coadsorption of  
459 antibiotic ciprofloxacin and Cu(II) and onto goethite surfaces. *Chem. Eng. J.*, **2015**, 269,  
460 113-120.
- 461 11. Martin, S.; Shchukarev, A.; Hanna, K.; Boily, J.-F. Kinetics and Mechanisms of  
462 Ciprofloxacin Oxidation on Hematite Surfaces. *Environ. Sci. Technol.*, **2015**, 49 (20),  
463 12197–12205.
- 464 12. Goyne, K. W.; Chorover, J.; Kubicki, J. D.; Zimmerman, A. R.; Brantley, S. L. Sorption  
465 of the antibiotic ofloxacin to mesoporous and nonporous alumina and silica. *J. Coll. Int.*  
466 *Sci.*, **2005**, 283, 160–170.
- 467 13. Paul, T.; Machesky, M.; Strathmann, T. Surface Complexation of the Zwitterionic  
468 Fluoroquinolone Antibiotic Ofloxacin to Nano-Anatase TiO<sub>2</sub> Photocatalyst Surfaces.  
469 *Environ. Sci. Technol.*, **2012**, 46 (21), 11896-11904.
- 470 14. Paul, T.; Liu, J.; Machesky, M.L.; Strathmann, T. Adsorption of zwitterionic  
471 fluoroquinolone antibacterials to goethite: A charge distribution-multisite complexation  
472 model. *J. Coll. Int. Sci.*, **2014**, 428, 63-72.
- 473 15. Pouliquen, H.; Le Bris, H. Sorption of oxolinic acid and oxytetracycline to marine  
474 sediments. *Chemosphere*, **1996**, 33 (5), 801-815.
- 475 16. Burns, R. G.; Burns, V. M. *Authigenic oxides. p. 875–914. In The Oceanic Lithosphere,*  
476 *Vol. 7, ed. Emiliani C., John Wiley & Sons, New York. 1981.*
- 477 17. Font, E.; Abrajevitch, A. Paleoenvironmental signature of the Deccan Phase-2  
478 eruptions. *Frontiers in Earth Sciences*, **2014**, 2(23), 1-5.
- 479 18. Selwyn, L. S.; Sirois, P. J.; Argyropoulos, V. The corrosion of excavated archaeological  
480 iron with details on weeping and akaganéite. *Stud. Conserv.*, **1999**, 44, 217-232.
- 481 19. Gismelseed, A.; Elzain, M.; Yousif, A.; Al Rawas, A.; Al-Omari, I. A.; Widatallah, H.;  
482 Rais, A. Identification of corrosion products due to seawater and fresh water. *Hyperfine*  
483 *Interact.*, **2004**, 156, 487–492.
- 484 20. Remazeilles, C.; Refait, P. On the formation of beta-FeOOH (akaganéite) in chloride-  
485 containing environments. *Corros. Sci.*, **2007**, 49, 844–857.
- 486 21. Bibi, I.; Singh, B.; Silvester, E. Akaganéite ( $\beta$ -FeOOH) precipitation in inland acid  
487 sulfate soils of south-western New South Wales (NSW), Australia. *Geochim.*  
488 *Cosmochim. Acta*, **2011**, 75, 6429-6438.
- 489 22. Cornell, R. M.; Schwertmann, U. *The iron oxides : structure, properties, reactions,*  
490 *occurrences, and uses*; Wiley-VCH: Weinheim, **2003**.
- 491 23. van der Zee, C.; Roberts, D.; Rancourt, D. G.; Slomp, C. P. Nanogoethite is the  
492 dominant reactive oxyhydroxide phase in lake and marine sediments. *Geology*, **2003**, 31  
493 (11), 993–996.

- 494 24. Gaboriaud, F.; Ehrhardt, J. J. Effects of different crystal faces on surface charge of  
495 colloidal goethite ( $\alpha$ -FeOOH) particles: an experimental and modeling study. *Geochim.*  
496 *Cosmochim. Acta*, **2003**, 67, 967-983.
- 497 25. Kozin, P. A.; Boily J.-F. Proton Binding and Ion Exchange at the Akaganéite/Water  
498 Interface. *J. Phys. Chem. C*, **2013**, 117, 6409–6419.
- 499 26. Tolls, J. Sorption of veterinary pharmaceuticals in soils: a review. *Environ. Sci.*  
500 *Technol.*, **2001**, 35 (17), 3397-3406.
- 501 27. Altmaier, M.; Metz, V.; Neck, V.; Müller, R.; Fanghänel, Th. Solid-liquid equilibria of  
502  $\text{Mg}(\text{OH})_{2(\text{cr})}$  and  $\text{Mg}_2(\text{OH})_3\text{Cl}\cdot 4\text{H}_2\text{O}_{(\text{cr})}$  in the system Mg-Na-H-OH-Cl- $\text{H}_2\text{O}$  at 25 °C.  
503 *Geochim. Cosmochim. Acta*, **2003**, 67, 3595-3601.
- 504 28. Hiemstra, T.; van Riemsdijk, W. H. A surface structural approach to ion adsorption: the  
505 charge distribution (CD) model. *J. Colloid Interf. Sci.*, **1996**, 179, 488-508.
- 506 29. Parkhurst, D. L.; Appelo, C. A. J. *User's guide to PHREEQC (Version 2) – a computer*  
507 *program for speciation, batch reaction, one-dimensional transport and inverse*  
508 *geochemical calculation*; Denver, Colorado, **1999**; p 312.
- 509 30. Kinniburgh, D. G.; Cooper, D. M. *PhreePlot: Creating graphical output with*  
510 *PHREEQC*. <http://www.phreeplot.org>. **2009**.
- 511 31. Powell, M. J. D. A method for minimizing a sum of squares of non-linear functions  
512 without calculating derivatives. *The Computer Journal*, **1965**, 7, 303-307.
- 513 32. Venema, P.; Hiemstra, T.; van Riemsdijk, W. H. Multisite Adsorption of Cadmium on  
514 Goethite. *J. Coll. Int. Sci.*, **1996**, 183, 515-527.
- 515 33. Lützenkirchen, J.; Marsac, R.; Kulik, D. A.; Payne, T. E.; Xue, Z.; Orsetti, S.;  
516 Haderlein, S. B. Treatment of multi-dentate surface complexes and diffuse layer  
517 implementation in various speciation codes. *Appl. Geochem.*, **2015**, 55, 128-137.
- 518 34. Johnson, B. B.; Sjöberg, S.; Persson, P. Surface Complexation of Mellitic Acid to  
519 Goethite: An Attenuated Total Reflection Fourier Transform Infrared Study. *Langmuir*,  
520 **2004**, 20, 823-828
- 521 35. Norén, K.; Persson, P. Adsorption of monocarboxylates at the water/goethite interface:  
522 The importance of hydrogen bonding. *Geochim. Cosmochim. Acta*, **2007**, 71, 5717-  
523 5730.
- 524 36. Boily, J.-F.; Lützenkirchen, J.; Blamès, O.; Beattie, J.; Sjöberg, S. Modeling proton  
525 binding at the goethite ( $\alpha$ -FeOOH)–water Interface. *Colloids Surf. A*, **2001**, 179, 11–27.
- 526 37. Song X.; Boily, J.-F. Surface Hydroxyl Identity and Reactivity in Akaganéite *J. Phys.*  
527 *Chem. C*, **2011**, 115, 17036-17045.
- 528 38. Song X.; Boily, J.-F. Competitive ligand exchange on akaganéite surfaces enriches bulk  
529 chloride loadings. *J. Coll. Int. Sci.*, **2012**, 376, 331-333.
- 530 39. IUPAC (2001) *IUPAC Stability Constants Database. Version 5.4, IUPAC and*  
531 *Academic Software* ([acadsoft@bcs.org.uk](mailto:acadsoft@bcs.org.uk)).
- 532 40. Neugebauer, U.; Szeghalmi, A.; Schmitt, M.; Kiefer, W.; Popp, J.; Holzgrabe, U.  
533 Vibrational spectroscopic characterization of fluoroquinolones. *Spectrochim. Acta, Part*  
534 *A*, **2005**, 61, 1505-1517.

- 535 41. Tarushi, A.; Christofis, P.; Psomas, G. Synthesis, characterization and interaction with  
536 DNA of mononuclear metal complexes with oxolinic acid. *Polyhedron*, **2007**, 26, 3963-  
537 3972.
- 538 42. Nakamoto, K. *Infrared and Raman Spectra of Inorganic and Coordination Compounds.*  
539 *Part B: Applications in Coordination, Organometallic, and Biorganic Chemistry, 5th*  
540 *ed.*; Wiley-Interscience: New York, **1997**.
- 541 43. Yost, E. C.; Tejedor-Tejedor, M. I.; Anderson, M. A. In Situ CIR-FTIR  
542 Characterization of Salicylate Complexes at the Goethite/Aqueous Solution Interface.  
543 *Environ. Sci. Technol.*, **1990**, 24, 822–828.
- 544 44. Cooper, E. M.; Vasudevan, D. Hydroxynaphthoic acid isomer sorption on goethite. *J.*  
545 *Col. Int. Sci.*, **2009**, 333, 85–96.
- 546 45. Rusch, B.; Hanna, K.; Humbert, B. Sorption and Transport of Salicylate in a Porous  
547 Heterogeneous Medium of Silica Quartz and Goethite. *Environ. Sci. Technol.*, **2010**, 44,  
548 2447–2453.
- 549 46. Hanna, K.; Boily, J.-F. Sorption of Two Naphthoic Acids to Goethite Surface under  
550 Flow through Conditions. *Environ. Sci. Technol.*, **2010**, 44, 8863–8869.
- 551 47. Boily, J.-F.; Persson, P.; Sjöberg, S. Benzenecarboxylate surface complexation at the  
552 goethite ( $\alpha$ -FeOOH)/water interface: II. Linking IR spectroscopic observations to  
553 mechanistic surface complexation models for phthalate, trimellitate, and pyromellitate.  
554 *Geochim. Cosmochim. Acta*, **2000**, 64 (20), 3453-3470.
- 555 48. Yates, J. T. Jr.; Madey, T. E. *Vibrational Spectroscopy of Molecules on Surfaces.*  
556 Plenum Press, **1987**.
- 557 49. Lai, H.-T.; Chien, Y.-H. ; Lin, J.-S. Long-term transformation of oxolinic acid in water  
558 from an eel pond. *Aquaculture*, **2008**, 275 (1-4), 96-101.
- 559 50. Psomas, G.; Tarushi, A.; Efthimiadou, E. K.; Sanakis, Y.; Raptopoulou, C. P.; Katsaros,  
560 N. Synthesis, structure and biological activity of copper(II) complexes with oxolinic  
561 acid. *J. Inorg. Biochem.*, **2006**, 100, 1764-1773.

562 **Table 1. Surface Complexation Model.**

					(101)	(001)	(210)
<b>Goethite<sup>a</sup></b>	log K	$\Delta z_0$	$\Delta z_1$	$\Delta z_2$	(Site/nm <sup>2</sup> )		
$\equiv\text{Fe}_3\text{O}^{-0.5} + \text{H}^+ \rightleftharpoons \equiv\text{Fe}_3\text{OH}^{+0.5}$	9.1	+1	0	0	3.03	3.34	-
$\equiv\text{Fe}_3\text{O}^{-0.5} + \text{H}^+ + \text{Cl}^- \rightleftharpoons \equiv\text{Fe}_3\text{OH}_2^{+0.5}\cdots\text{Cl}^-$	8.1	+1	0	-1			
$\equiv\text{Fe}_3\text{O}^{-0.5} + \text{Na}^+ \rightleftharpoons \equiv\text{Fe}_3\text{OH}^{-0.5}\cdots\text{Na}^+$	-1	0	0	+1			
$\equiv\text{FeOH}^{-0.5} + \text{H}^+ \rightleftharpoons \equiv\text{FeOH}_2^{+0.5}$	9.1	+1	0	0	3.03	3.34	7.4
$\equiv\text{FeOH}^{-0.5} + \text{H}^+ + \text{Cl}^- \rightleftharpoons \equiv\text{FeOH}_2^{+0.5}\cdots\text{Cl}^-$	8.1	+1	0	-1			
$\equiv\text{FeOH}^{-0.5} + \text{Na}^+ \rightleftharpoons \equiv\text{FeOH}^{-0.5}\cdots\text{Na}^+$	-1	0	0	+1			
$2 \text{H}^+ + 2 \equiv\text{FeOH}^{-0.5} + \text{OA}^- \rightleftharpoons (\equiv\text{FeOH}_2)_2^+\cdots(\text{OA})^-$	20.3±0.1 <sup>c</sup>	+2	-1	0			
$2 \text{H}^+ + 2 \equiv\text{FeOH}^{-0.5} + \text{OA}^- \rightleftharpoons (\equiv\text{Fe})_2(\text{OA})^0 + 2 \text{H}_2\text{O}$	20.3±0.1 <sup>c</sup>	+1	0	0			
<b>Akaganéite<sup>b</sup></b>	log K	$\Delta z_0$	$\Delta z_1$	$\Delta z_2$	(100)/(001)	(010)	
					(Site/nm <sup>2</sup> )		
$\equiv\text{FeO}^{-1.5} + \text{H}^+ \rightleftharpoons \equiv\text{FeOH}^{-0.5}$	20.0	+1	0	0	3.09		-
$\equiv\text{FeO}^{-1.5} + 2 \text{H}^+ \rightleftharpoons \equiv\text{FeOH}_2^{+0.5}$	28.2	+2	0	0			
$\equiv\text{FeO}^{-1.5} + 2 \text{H}^+ + \text{Cl}^- \rightleftharpoons \equiv\text{FeOH}_2^{+0.5}\cdots\text{Cl}^-$	29.2	+2	0	-1			
$\equiv\text{FeO}^{-1.5} + \text{H}^+ + \text{Na}^+ \rightleftharpoons \equiv\text{FeOH}^{-0.5}\cdots\text{Na}^+$	21.0	+1	0	+1			
$2 \text{H}^+ + 2 \equiv\text{FeOH}^{-0.5} + \text{OA}^- \rightleftharpoons (\equiv\text{FeOH}_2)_2^+\cdots(\text{OA})^-$	21.5±0.1 <sup>c</sup>	+2	-1	0			
$2 \text{H}^+ + 2 \equiv\text{FeOH}^{-0.5} + \text{OA}^- \rightleftharpoons (\equiv\text{Fe})_2(\text{OA})^0 + 2 \text{H}_2\text{O}$	21.5±0.1 <sup>c</sup>	+1	0	0			
$\equiv\text{Fe}_3\text{O}_I^{-0.5} + \text{H}^+ \rightleftharpoons \equiv\text{Fe}_3\text{O}_I\text{H}^{+0.5}$	11.7	+1	0	0	6.18		-
$\equiv\text{Fe}_3\text{O}_{II}^{-0.5} + \text{H}^+ \rightleftharpoons \equiv\text{Fe}_3\text{O}_{II}\text{H}^{+0.5}$	2.6	+1	0	0	3.09		-
$\equiv\text{Fe}_2\text{O}^{-1} + \text{H}^+ \rightleftharpoons \equiv\text{Fe}_2\text{OH}^0$	10.8	+1	0	0	3.09		7.06
$\equiv\text{Fe}_2\text{O}^{-1} + 2 \text{H}^+ \rightleftharpoons \equiv\text{Fe}_2\text{OH}_2^+$	17.9	+2	0	0			
$\equiv\text{Fe}(\text{OH}_2)(\text{OH})^0 + \text{H}^+ \rightleftharpoons \equiv\text{Fe}(\text{OH}_2)_2^+$	8.2	+1	0	0	-		3.53
$\text{H}^+ + \equiv\text{Fe}(\text{OH}_2)(\text{OH})^0 + \text{OA}^- \rightleftharpoons \equiv\text{Fe}(\text{OA})^0 + 2 \text{H}_2\text{O}$	13.9±0.1	0	0	0			

563

564 <sup>a</sup>TPM with  $C_1=2.3 \text{ F/m}^2$ ,  $C_2=1.07 \text{ F/m}^2$ , surface area: 95 m<sup>2</sup>/g; 63% of (101), 27% of (001) and 10% of  
 565 (210). The HB surface complex is considered to form only at the (101) and (001) planes while the MB  
 566 occurs at all planes.

567 <sup>b</sup>TPM with  $C_1=2.3 \text{ F/m}^2$ ,  $C_2=1.6 \text{ F/m}^2$ , surface area: 239 m<sup>2</sup>/g; 95% of (100)/(001) and 5% of (010).

568 <sup>c</sup>Constant and uncertainty obtained with Phreeplot, corrected according to Venema et al.<sup>32</sup>.

569

570



571 **Figure captions**

572 **Figure 1.** OA removal from solution (molecule / nm<sup>2</sup>) for [OA]<sub>tot</sub>=10 μM on 50 m<sup>2</sup>/L goethite  
573 (a) and akaganéite (b) versus p<sub>Hc</sub> (= -log [H<sup>+</sup>]) at different NaCl concentrations. OA removal  
574 from solution (molecule / nm<sup>2</sup>) for [OA]<sub>tot</sub>=50 μM on 50 m<sup>2</sup>/L goethite (c) and akaganéite (d)  
575 in 100 mM NaCl versus pH (= -log [H<sup>+</sup>]). The corresponding percentage of OA removal at the  
576 plateau is also given. Lines are modeling results, where OA speciation on the solid phase is  
577 shown for [NaCl]=100 mM (legend in a/c and b/d is the same for the model). MB and HB  
578 refer to metal- and hydrogen-bonded OA, respectively.

579

580 **Figure 2.** (a) Experimental and calculated OA solubility in 100 mM NaCl versus p<sub>Hc</sub>. The  
581 equilibrium OA concentration ([OA]<sub>aq</sub>) in contact with goethite and akaganéite for [OA]<sub>tot</sub>=10  
582 and 50 μM in 100 mM NaCl are also plotted with modeling results. (b) OA-goethite and OA-  
583 akaganéite sorption isotherms at p<sub>Hc</sub>=6.1±0.1 and 100 mM NaCl. The results of the models  
584 for goethite and akaganéite and OA speciation at the akaganéite surface are also shown. OA  
585 solubility limit at p<sub>Hc</sub>=6.1 is shown as a vertical dashed line.

586

587 **Figure 3.** ATR-FTIR spectra of OA adsorbed to goethite (a) and akaganéite (b) for various  
588 p<sub>Hc</sub> (-log [H<sup>+</sup>]) in 100 mM NaCl for [OA]<sub>tot</sub>=50 μM (except for goethite at p<sub>Hc</sub>=5.20 and  
589 6.39, where [OA]<sub>tot</sub>=20 μM). ATR-FTIR spectra of dissolved OA in 1 M NaOH (OA<sub>(aq)</sub>) and  
590 protonated OA under its solid form (OA<sub>(s)</sub>) are shown as references.

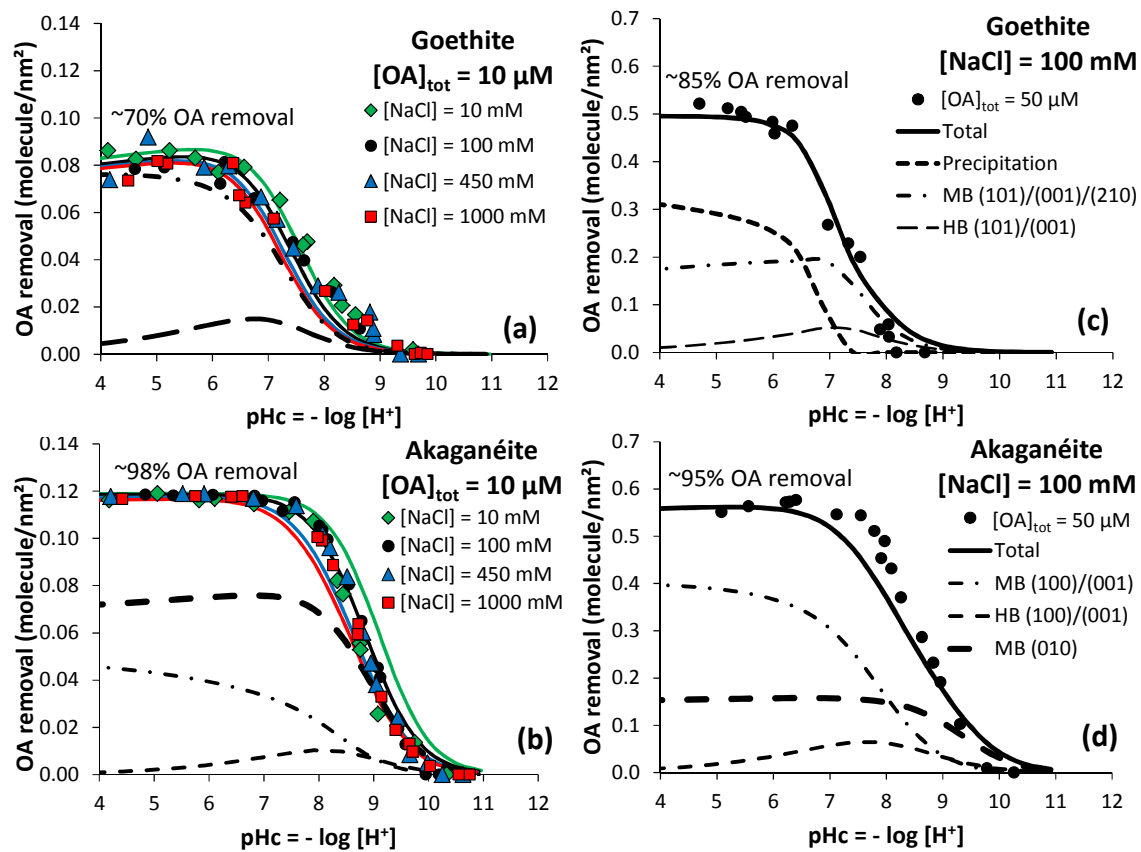
591

592 **Figure 4.** Possible surface complexes of OA at goethite and akaganéite surfaces. The  
593 electronic structure of OA is proposed according to previous studies on metal-OA  
594 compounds.<sup>41,50</sup>

595

Figure 1

596



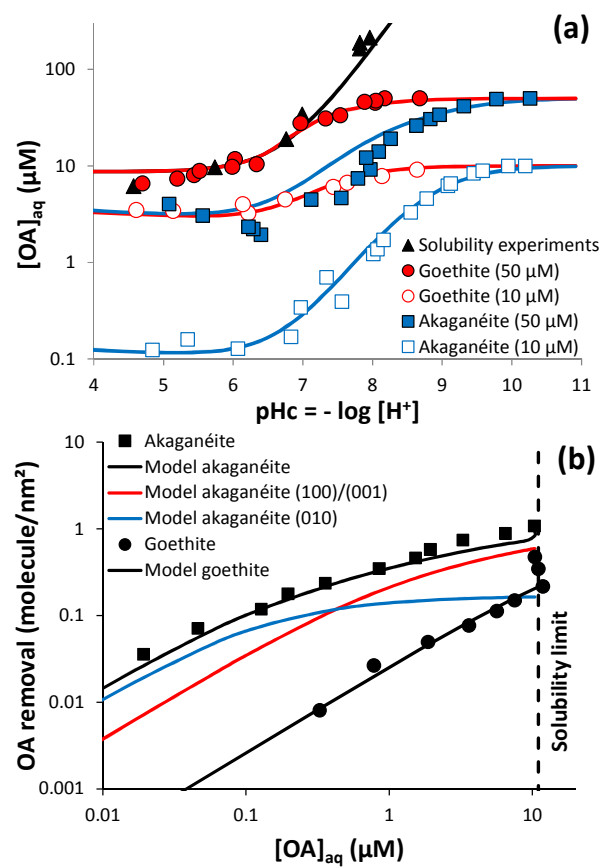
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Figure 2



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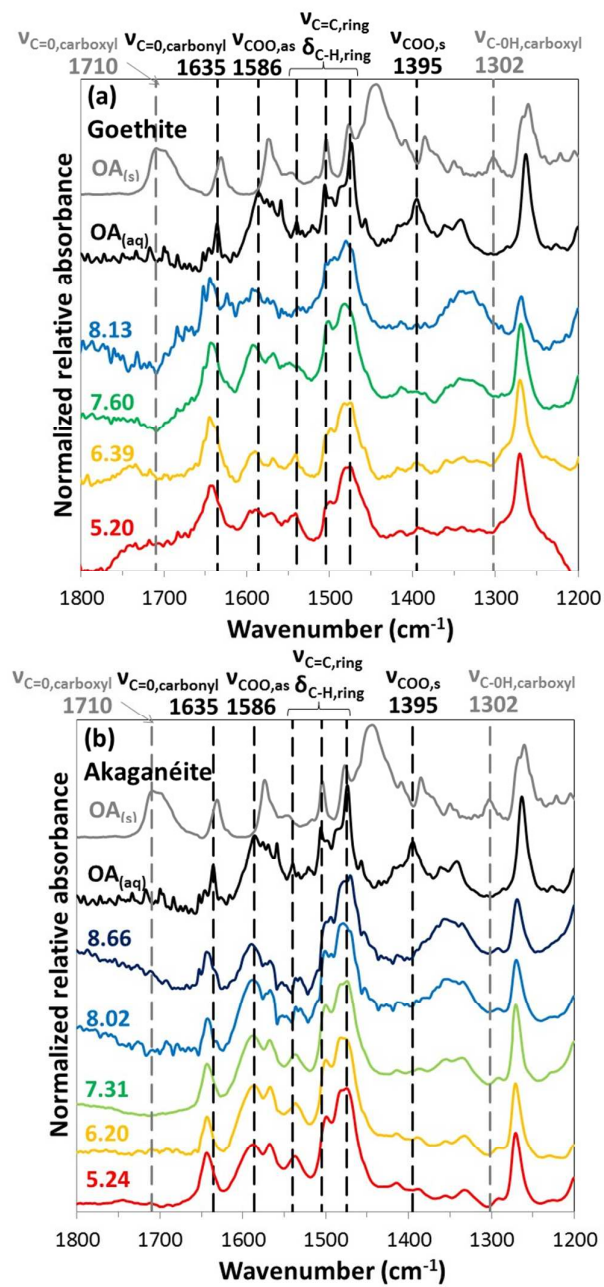
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Figure 3

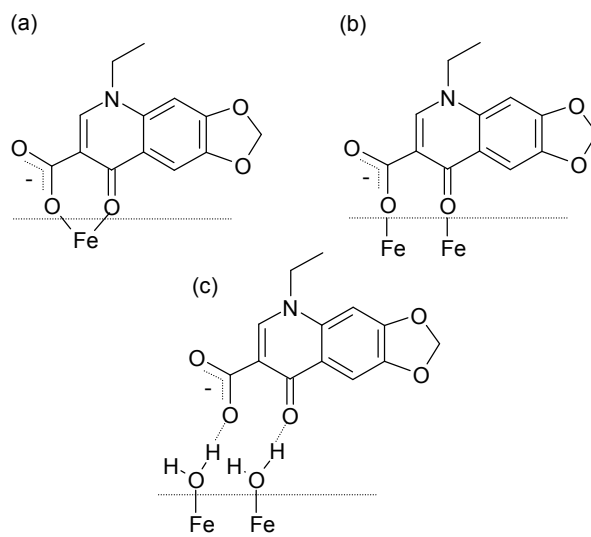


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Figure 4

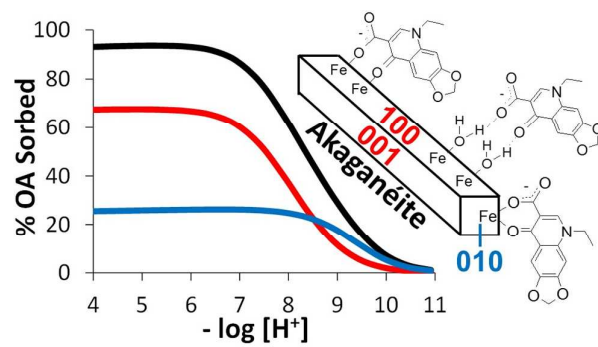


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## TOC/abstract



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