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Oxolinic Acid Binding at Goethite and Akaganéite Surfaces: Experimental Study and Modeling

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24 **Abstract**

25 Oxolinic acid (OA) is a widely used quinolone antibiotic in aquaculture. In this study, its
26 interactions with synthetic goethite (α -FeOOH) and akaganéite (β -FeOOH) particle surfaces
27 were monitored to understand the potential fate of OA in marine sediments where these
28 phases occur. Batch sorption experiments, liquid chromatography (LC) analyses of
29 supernatants, Attenuated Total Reflectance-Fourier Transform InfraRed (ATR-FTIR)
30 spectroscopy and multi-site complexation (MUSIC) modeling were used to monitor OA
31 binding at these particle surfaces. Both LC and ATR-FTIR showed that adsorption did not
32 degrade OA, and that OA adsorption was largely unaffected by NaCl concentrations (10-1000
33 mM). This was explained further by ATR-FTIR suggesting the formation of metal-bonded
34 complexes at circumneutral to low pHc ($= -\log [H^+]$) and with a strongly hydrogen-bonded
35 complex at high pHc. The stronger OA binding to akaganéite can be explained both by the
36 higher isoelectric point/point-of-zero charge (9.6-10) of this mineral than of goethite (9.1-9.4),
37 and an additional OA surface complexation mechanism at the (010) plane. Geminal sites
38 ($\equiv Fe(OH_2)_2^+$) at this plane could be especially reactive for metal-bonded complexes, as they
39 facilitate a mononuclear six-membered chelate complex via the displacement of two
40 hydroxo/aquo groups at the equatorial plane of a single Fe octahedron. Collectively, these
41 findings revealed that Fe-oxyhydroxides may strongly contribute to the fate and transport of
42 OA-type antibacterial agents in marine sediments and waters.

43

1. Introduction

Oxolinic Acid (OA) belongs to a family of quinolone antibiotics that is commonly used in aquaculture as a prophylactic, or as a chemotherapy agent.^{1,2} Unabsorbed or excreted OA released into waters and/or accumulated in marine and freshwater sediments poses potential risks to public health and to aquatic life.³ Due to its frequent use in fish farming, concentrations of OA can be of the order of 0.01–2.50 ppm in fresh or saline surface waters, 0.01–2.31 ppm in bottom waters, and even 1.81–426.31 ppm in pond sediments.^{1,4,5} Moreover, Greek fish farms release 50 kg of OA annually to the Mediterranean Sea.⁴

In marine or freshwater sediments, soils and aquatic environments, the transport and mobility of quinolone antibiotics are strongly related to reactive oxide minerals through adsorption and redox reactions.⁶⁻¹⁴ In a study on oxolinic acid sorption on marine sediments Pouliquen and Le Bris¹⁵ concluded that while OA was not transformed in the sediments by abiotic or biotic processes, it can be sorbed to sediments containing reactive mineral surfaces. Among these reactive surfaces, iron oxyhydroxides, including goethite (α -FeOOH) and akaganéite (β -FeOOH), are ubiquitous in marine environments.^{16,17} Akaganéite is a chlorine-bearing iron oxyhydroxide with a hollandite-type (tunnel-like) structure containing Cl^- ions residing in 4×4 nm wide tunnels running along the length of acicular nano-sized particles. This phase can be generally found in environments rich in Fe(II) and Cl^- , and as a natural corrosion product in high salinity environmental settings.¹⁸⁻²¹ Goethite is the most thermodynamically stable iron oxyhydroxide mineral and is reported to form through the transformation of metastable phases such as ferrihydrite, schwertmannite and akaganéite.²² Goethite was also shown to be the most common diagenetic iron oxyhydroxide in both marine and lake sediments.²³

Although OA has been extensively used in industrial fish farming, and thus is frequently detected in surrounding waters or sediments, little is known about its molecular

interactions with suspended particles or mineral constituents in sediments, which is necessary to accurately assess its fate in aquatic environments. In this study, interactions of OA with goethite and akaganéite particles were studied by batch experiments, surface complexation modeling, and vibration spectroscopic investigations. Effects of pHc ($4 < -\log [H^+] < 11$), salinity (10-1000 mM NaCl) and surface loading on OA binding mechanisms were resolved, and implications for the fate of OA in marine environments were proposed.

2. Materials and methods

2.1. Chemicals

All chemicals used were of pro-analytical quality or better and were purchased from Sigma-Aldrich. An OA (purity > 99%) stock solution was prepared by dissolving 30 mg (115 μ moles) OA in 20 mL of 1 M NaOH, then diluted to 1 L with ultrapure water.

2.2. Synthesis and characterization of goethite and akaganéite particles

Goethite and akaganéite particles were synthesized as described in previous studies.^{24,25} The synthesis procedure is given in detail in the supporting information (SI). The obtained particles were characterized by particle size and B.E.T. specific surface area $N_2(g)$ adsorption measurements, chemical analysis, transmission electron microscopy (TEM) and X-ray diffraction (XRD). The B.E.T. specific surface area of the synthetic goethite and akaganéite were 95 and 239 m^2/g , respectively. Other characteristic properties of these solids are summarized in the SI.

2.3. Batch experiments and liquid chromatography analysis

Because high background electrolyte solutions can affect pH measurements,²⁷ the pH electrode was calibrated to measure the molarity of the proton ($-\log [H^+]$, noted pHc) using

solutions of known $[H^+]$ (10^{-5} - 10^{-3} M) at different $[NaCl]$, and the results are provided in this form in this study.

OA solubility experiments (undersaturation direction) were conducted by suspending solid OA (~3-5 mg) in 10 mL of solution in 100 mM NaCl as a function of pHc. The suspensions were equilibrated for 24 h, after which the supernatants were filtered (0.2 μ m) and OA concentrations were measured with high pressure liquid-chromatography coupled with UV detection (HPLC-UV). As discussed later, OA solubility is relatively low (~10 μ M) at acidic pHc, in agreement with previous studies.²⁶

An akaganéite (stock) suspension was prepared as previously reported.²⁵ Aliquots of the original stock suspension (21 g/L, 5023 m²/L) were used to prepare four suspensions (2.1 g/L, 502.3 m²/L) in different concentrations of NaCl (10, 100, 450 or 1000 mM). These suspensions were equilibrated for over 30 days to ensure complete ion exchange between bulk akaganéite and solution ions, as previously explained by Kozin and Boily.²⁵ Goethite suspensions were prepared at the chosen ionic strength just before adding OA.

Batch adsorption experiments were performed under conditions of varied reaction time, OA concentration, pHc and ionic strength. Kinetic studies were conducted for pHc = 6, $[NaCl] = 100$ mM, $[OA]_{tot} = 50$ μ M with an equivalent surface area of solids, i.e. 50 m²/L of goethite or akaganéite. All batch experiments were performed under an atmosphere of N₂(g) to purge dissolved CO₂ from the aqueous solutions. Briefly, solutions of 1 to 100 μ M OA were mixed with goethite or akaganéite (50 m²/L) at the desired concentration of NaCl (10, 100, 450 or 1000 mM). pHc was thereafter adjusted to the desired value by addition of 0.1 M HCl solutions (the initial pHc value was about 10). In another set of experiments, adsorption isotherms were collected at pHc = 6.1 $\pm$ 0.1, for $[OA]_{tot}$ concentrations in the 1-50 μ M range for goethite, and the 3-100 μ M range for akaganéite. All suspensions were then equilibrated for 24 h and the supernatants filtered (0.2 μ m) prior to soluble OA determination by HPLC-

UV. Desorption tests ($\text{pHc}=11$) were conducted under various experimental conditions as a means to verify the mass balance of OA in the systems under study.

Soluble OA concentrations were determined by a Waters 600 controller high performance liquid chromatography (HPLC) system equipped with an auto sampler (Waters 717 plus), a C18 column (250 mm $\times$ 4.6 mm i.d., 5 μm) and a UV detector (275 nm, Waters 2489). The mobile phase was a mixture of water/acetonitrile (60:40 v/v) containing 0.1% of formic acid. The flow rate of the mobile phase was set at 1 mL/min in the isocratic mode. Finally, the complete experimental solubility and sorption datasets are provided in SI (Tables S1, S2 and S3).

2.4. ATR-FTIR spectroscopy

Attenuated total reflectance-Fourier transform infrared (ATR-FTIR) spectra were recorded between in the 780-4000 cm^{-1} region on an IS50 Nicolet spectrometer equipped with a KBr beamsplitter and a liquid Nitrogen cooled MCT (Mercury Cadmium Telluride) detector. A nine-reflection diamond ATR accessory (DurasamplIRTM, SensIR Technologies) was used for acquiring spectra of wet samples. The resolution of the single beam spectra was 4 cm^{-1} .

Three series of experiments were conducted with akaganéite for: (i) $5 < \text{pH} < 7$ with $[\text{OA}]_{\text{tot}}=50 \mu\text{M}$ and $[\text{NaCl}]=100 \text{ mM}$, (ii) $10 \leq [\text{NaCl}] \leq 1000 \text{ mM}$ with $[\text{OA}]_{\text{tot}}=50 \mu\text{M}$ and $\text{pHc}=6$ and (iii) $20 \leq [\text{OA}]_{\text{tot}} \leq 50 \mu\text{M}$ with $\text{pHc}=6$ and $[\text{NaCl}]=100 \text{ mM}$. For goethite, the same experiments were conducted except for $\text{pHc} < 7$, where $[\text{OA}]_{\text{tot}}=20 \mu\text{M}$ because OA partially precipitated at 50 μM (see below). Prior to ATR-FTIR analysis, tubes from batch sorption experiments were centrifuged at 10,000 g for 30 min. Wet mineral pastes were directly and uniformly applied to the diamond ATR crystal then covered with a lid of a flow-through cell to prevent the evaporation of water. ATR-FTIR spectra were then recorded

immediately. Reference spectra include those for pure water, the filtered (0.2 μm) supernatant from each batch adsorption experiment, as well as $\text{OA}_{(\text{s})}$ in its solid form but covered with a drop of water to apply it more uniformly. A reference spectrum of aqueous OA was also acquired from a 10 mM OA (2.5 g/L) solution in 1 M NaOH. This solution ensured a sufficiently high concentration of $\text{OA}^-_{(\text{aq})}$ for ATR-FTIR analysis. Otherwise, OA was not detected in any of the supernatants given its low solubility. All spectra are shown for the fingerprint 1200–1800 cm^{-1} region only. Additional ATR-FTIR measurements of OA in D_2O $\text{OA}^-_{(\text{aq})}$, OA-goethite (pD = 7.81) and OA-akanagite (pD = 8.51) in D_2O were also carried out to resolve bands arising from moieties that are directly affected by exchange or hydrogen-bonding with water. Sample preparation is detailed further in the Supporting Information (SI).

2.5. Surface complexation modeling

Predictions of sorption vs pHc were made using the multi-site complexation (MUSIC) model approach.²⁸ The geochemical speciation code PHREEQC (version 2)²⁹ was used for surface complexation calculations. PHREEQC includes the three plane model (TPM). Charges of the adsorbates were distributed among the 0 (H^+ , metal-bonded OA), 1 (hydrogen-bonded OA), and 2 (Na^+ , Cl^-) planes of the TPM. The TPM requires two capacitances for the 0- and the 1-plane, C_1 and C_2 , respectively (see SI for more details). PhreePlot³⁰ was used to estimate parameters for OA sorption to goethite and akaganite. Using a modified Marquardt-Levenberg procedure,³¹ PhreePlot provides a statistical uncertainty for the estimated parameters, which are reported in this study. For goethite or akaganite, different types of surface sites can exist depending on the proportion of crystal planes, which should have different binding affinities. All information concerning the dominant crystal planes of particles and corresponding site densities are given in the SI. Charge densities and corresponding electrostatic interactions at particle termination (i.e. (021) plane of goethite and

(010) plane of akaganéite) were computed individually, as well as formation constants of all surface species are reported in Table 1. Surface complexation constants reported for OA were corrected according to Venema *et al.*³² (see also³³).

3. Results and Discussion

3.1. Batch experiments

Kinetic batch adsorption experiments conducted over a 7-day period showed that equilibrium was achieved within 2 h of reaction time (Figure S1). Desorption tests also showed that OA was removed only by adsorption or precipitation, and that breakdown by, for example, oxidation, did not occur under the experimental conditions of this study.

Adsorption curves for 10 μM OA solutions reacted in 50 m^2/L mineral suspensions (Fig. 1a,b) revealed greater OA loadings on akaganéite than on goethite. At $\text{pH}_c < 6.5$, only ~70% of OA sorbs onto goethite (Fig. 1a, Table S1) whereas an almost complete uptake was achieved on akaganéite at $\text{pH}_c=7.5$ (~98%; Fig. 1b, Table S2). OA adsorption was greatest under acidic to circumneutral pH, and lowest under alkaline conditions, as typically encountered for anionic ligands. Furthermore, an insignificant effect of NaCl concentration was detected, so OA binding should predominantly occur as strong metal-bonded complexes with surface Fe sites and/or directly hydrogen-bonded complexes with surface hydroxo groups. Resistance of the latter to variations in ionic strength was notably demonstrated by Johnson *et al.*³⁴. Weaker outer-sphere complexes should, on the other hand, be more affected by variations in ionic strength.³⁵ This observation is consistent with the adsorption of flumequine (analogous structure of OA) to goethite where the effect of ionic strength was also negligible ($1 < [\text{NaCl}] < 100 \text{ mM}$).¹⁴ Our explanation for the greater OA loadings achieved in akaganéite is two-fold. First, we note that while reactive site density of these two minerals under study are highly comparable (see details in SI), the point-of-zero charge of akaganéite

(9.6-10)²⁵ is higher than that of goethite (9.1-9.4)^{24,36}. Second, geminal $\equiv\text{Fe}(\text{OH})_2^+$ groups of the (010) plane of akaganéite are expected to be strongly reactive sites for binding anionic species.^{25,37,38} The site density of these groups is estimated at 0.18 site/nm² (i.e. 5% $\times$ 3.53 site/nm²), which may favor an additional OA removal mechanism, as will be suggested further by modeling (Fig. 1b and 1d).

In case of goethite ($\text{pHc} < 7$), the plateau is found higher for $[\text{OA}]_{\text{tot}}=50 \mu\text{M}$ (~85%) than for $[\text{OA}]_{\text{tot}}=10 \mu\text{M}$ (~70%) (Fig. 1a and 1c, Table S1), i.e. OA loading is more than 5 times larger at 50 μM than at 10 μM , which may suggest an additional removal mechanism such as surface or bulk precipitation of OA in this pHc range. The latter possibility was tested by conducting solubility experiments of OA at different pHc values (Fig. 2a, Table S3). Additionally, the OA solubility was calculated with PHREEQC using the acidity constant (K_a) value from the IUPAC stability constant database³⁹ and the solubility constant (K_s) fitted to our data with Phreeplot:

The experimental solubility values are in good agreement with both calculated (Fig. 2a) and literature data (e.g. 23 μM at $\text{pH} = 7$).²⁶ The equilibrium OA concentrations ($[\text{OA}]_{\text{aq}}$) in the presence of goethite or akaganéite at $[\text{OA}]_{\text{tot}}=50 \mu\text{M}$ are also plotted in Figure 2a. In contact with akaganéite, $[\text{OA}]_{\text{aq}}$ is below the solubility limit for all pHc tested, ruling out the bulk precipitation in this experimental set. $[\text{OA}]_{\text{aq}}$ values in goethite are however closer to the solubility curve at $\text{pHc} < 7.5$, a result possibly explaining the larger loadings achieved at $[\text{OA}]_{\text{tot}}=50 \mu\text{M}$. Accordingly, $[\text{OA}]_{\text{aq}}$ values measured at low ligand loadings ($[\text{OA}]_{\text{tot}}=10 \mu\text{M}$) are below the OA solubility limit (Fig. 2a).

Adsorption isotherms (Fig. 2b) add further validation to solubility data where a sharp rise in OA removal with goethite occurred at $\text{pHc}=6.1 \pm 0.1$ and $[\text{OA}]_{\text{aq}} \approx 10 \mu\text{M}$,

corresponding to the solubility limit of OA. The OA adsorption isotherm for akaganéite does not, on the other hand, reach this solubility limit, but rather reaches a maximal loading corresponding to $Q_{\max}=1.14$ OA/nm². This value was obtained by modeling OA-akaganéite data with the Langmuir adsorption isotherm:

$$[OA]_{\text{sorbed}} = Q_{\max} \frac{K_L \times [OA]_{\text{aq}}}{1 + K_L \times [OA]_{\text{aq}}} \quad (3)$$

with an adsorption constant of $K_L = 0.53 \mu\text{M}^{-1}$ (see the fit with the Langmuir equation in Fig. S2). As this $Q_{\max}$ value is about one-third our estimated reactive site density at FeOOH surfaces (*i.e.* ~ 3 -OH/nm²), OA surface complexes binding may realistically involve two or three -OH sites. The molecular-level nature of these interactions will however be addressed in greater detail in the following section.

3.2. ATR-FTIR Spectroscopic investigations

3.2.1. $OA_{(s)}$ and $OA_{(aq)}^-$ in 1 M NaOH

The ATR-FTIR spectra of solid and soluble OA (Fig. 3) were used as a basis for assigning surface-bound OA species. Owing to the low solubility of OA, band assignments in the soluble species were made from a 1 M NaOH solution (Fig. 3).

The absorption band of $OA_{(s)}$ and $OA_{(aq)}$ at 1040 cm⁻¹ (see Figure S3) corresponds to the stretching modes of the ether group ($\nu_{\text{C-O-C}}$). The FTIR spectra of $OA_{(s)}$ and $OA_{(aq)}$ exhibit three bands with maximum absorbance between 1470 and 1510 cm⁻¹. These bands are assigned to C=C stretches ($\nu_{\text{C=C,ring}}$) and C-H bends ($\delta_{\text{C-H,ring}}$) in the aromatic and quinolinone rings.⁴⁰ The intense band of $OA_{(s)}$ at 1440 cm⁻¹, which disappears in the spectrum of $OA_{(aq)}$, may be attributed to perturbation of $\nu_{\text{C=C,ring}}$ or $\delta_{\text{C-H,ring}}$ due to intermolecular interaction in $OA_{(s)}$. The major difference between ATR-FTIR spectra of $OA_{(s)}$ and $OA_{(aq)}$ lies in the

disappearance of two bands at 1705 and 1302 cm^{-1} from the deprotonation of the carboxyl groups, *i.e.* disappearance of C=O and C-OH stretching modes of the carboxylic group. This deprotonation gives, in turn, rise to the 1395 and 1586 cm^{-1} bands, which are respectively assigned to symmetric ($\nu_{\text{COO},s}$) and asymmetric ($\nu_{\text{COO},as}$) stretching modes of the carboxylate, and where $\Delta\nu = \nu_{\text{COO},as} - \nu_{\text{COO},s} = 191 \text{ cm}^{-1}$. These values are, moreover, consistent with previous studies on (fluoro)quinolones, ciprofloxacin (CIP) assigning $\nu_{\text{COO},s}$ to a band at 1360-1380 cm^{-1} ,⁶⁻¹¹ and on ofloxacin (OFX) with a similar band at 1340 cm^{-1} .¹²⁻¹⁴ Our spectrum for $\text{OA}_{(\text{aq})}$ also showed a distinct band at 1340 cm^{-1} , but for reasons given later, it is not assigned to $\nu_{\text{COO},s}$. The absorption band at 1635 cm^{-1} is assigned to the stretching of the quinolinone ring carbonyl group ($\nu_{\text{C=O,carbonyl}}$), in accordance with previous spectroscopic studies dedicated to (fluoro)quinolones.^{6-14,41}

3.2.2. *OA-goethite*

Figure 3a shows the effect of pHc on the ATR-FTIR spectra of 20 μM OA (for pHc < 7) or 50 μM OA (for pHc > 7) reacted with 50 m^2/L goethite in 100 mM NaCl. As shown in Figure S3a, the characteristic features of both $\text{OA}_{(\text{s})}$ and $\text{OA}_{(\text{aq})}$ were observed for OA-goethite samples at pHc = 6.12 for $[\text{OA}]_{\text{tot}} = 50 \mu\text{M}$. Only the characteristic of $\text{OA}_{(\text{aq})}$ appear for pHc $\approx$ 6 with $[\text{OA}]_{\text{tot}} \leq 30 \mu\text{M}$ and for pH > 7 with $[\text{OA}]_{\text{tot}} = 50 \mu\text{M}$ (Fig. 2, Fig. S3), in agreement with our batch results. Especially, the absence of band at 1705 cm^{-1} suggests the presence of deprotonated carboxylate groups.

At pHc = 5.20 and 6.39 ($[\text{OA}]_{\text{tot}} = 20 \mu\text{M}$), the band at 1395 cm^{-1} for $\text{OA}_{(\text{aq})}$ almost completely disappears and is replaced by a band centered at 1340 cm^{-1} . This further supports the assignment of $\nu_{\text{COO},s}$ to the band at 1395 cm^{-1} in $\text{OA}_{(\text{aq})}$. The $\nu_{\text{COO},as}$ band might be slightly blue-shifted (1590 cm^{-1}), which gives $\Delta\nu \approx 250 \text{ cm}^{-1}$, a value suggesting that only one oxygen of the carboxylate binds to a surface Fe.⁴² A blue shift of $\nu_{\text{C=O,carbonyl}}$ (1642 cm^{-1}) compared to

OA_(aq) or OA_(s), also suggests the involvement of the keto group in the surface complex. A similar shift from protonated OA_(s) to Fe(OA)_{3(s)} was previously reported, where OA forms a bidentate deprotonated complex with Fe³⁺ through the keto group and one carboxylate oxygen.⁴¹ The effects of NaCl concentration were also investigated for 20 μM OA and 50 m²/L goethite at pH_c ≈ 6 (Fig. S3b). The weak ionic strength dependence on OA loadings and ATR-FTIR spectra point to the predominance of OA species that are inner-spherically bound to surface Fe sites (*i.e.* metal-bonded complex; MB), involving one oxygen of the carboxylate and the oxygen of the keto group of OA. The involvement of both the carboxylate and keto groups was proposed in previous studies dedicated to other quinolones binding to hydrous ferric oxides⁶, magnetite⁹, hematite¹¹ and goethite^{10,14} or observed for metal cation-OA complexes in water.⁴¹ The involvement of both the carboxylate and the phenolate of salicylate and hydroxynaphthoate has also been evidenced at goethite surfaces.⁴³⁻⁴⁶ We cannot directly conclude from the ATR-FTIR spectra whether this complex involves one (Fig. 4a) or two surface Fe atoms (Fig. 4b). Ideally, the steric constraints at the (101)/(001) as well as the (210) planes (Pnma group) should promote bridging between two Fe atoms separated by 3 Å from one another (Fig. 4b). Moreover, defects ought to expose additional possibilities for mononuclear bidentate binding (*cf.* six-membered chelate complexation, Fig. 4a), namely at equatorial edges of a Fe octahedron.⁴⁷

At pH_c = 7.60 and 8.13 ([OA]_{tot}=50μM), all bands remain unaffected compared to pH_c = 5.20 and 6.39 ([OA]_{tot}=20μM), except the band at 1340 cm⁻¹, which broadens and whose full width at half maximum considerably increases with pH_c. This band broadening is taken as an indication of interactions of carboxylic groups with surface hydroxo groups with various hydrogen-bond strengths (*cf.* inhomogeneous band broadening⁴⁸). An ATR-FTIR spectrum of OA-goethite recorded in D₂O for pD = 7.81 ([OA]_{tot}=50μM) reveals that the 1340 cm⁻¹ band is the only one that is red-shifted in relation to the counterpart spectrum in

H₂O (Fig. S4), thus suggesting close interactions between carboxyl groups and O-D/H moieties. These observations could thus point to a hydrogen-bonded (HB) complex involving a direct hydrogen bond between carboxyl and keto moieties and surface (hydr)oxo groups. This ionic strength independent complex thus contrasts with outer-sphere complexes involving hydration waters.³⁴ The increased intensity of the band at 1340 cm⁻¹ suggests that a strong form of hydrogen bonding is favored at high pHc whereas metal-bonded complexation is favored at low pHc. This interpretation is in line with, for example, phthalate or hydroxynaphthoate sorption to goethite^{44,46,47} where HB complexes are involved with two –OH sites at the (101)/(001) planes (Fig. 4c).

3.2.3. OA-akaganéite

Figure 3b shows the effect of pHc on ATR-FTIR spectra of 50 μM OA interacting with 50 m²/L akaganéite in 100 mM NaCl. In agreement with our batch sorption results (Fig. 2), the characteristic features of OA_(s) are absent and therefore spectra must be attributed to the surface complexed OA only. This is further confirmed by the analysis of OA-akaganéite samples at lower [OA]_{tot} (down to 20 μM) at pHc ≈ 6 in 100 mM NaCl (Fig. S3c). The spectra recorded at pHc = 5.24 and 6.20 are similar to those of OA-goethite when the amount of precipitated OA is negligible, namely $\nu_{\text{C=O,carbonyl}} = 1642 \text{ cm}^{-1}$ and $\nu_{\text{COO,s}} = 1330 \text{ cm}^{-1}$. The intense band observed for OA_(aq) at 1395 cm⁻¹ disappears in the spectra recorded for OA-akaganéite. The band of $\nu_{\text{COO,as}}$ (1590 cm⁻¹) is also slightly blue-shifted compared to OA_(aq). As for goethite, no significant effect of NaCl concentration on OA sorption to akaganéite can be observed for pHc ≈ 6 (Fig. 1b), in accordance with the ATR-FTIR spectra (Fig. S3d). Because these experimental conditions correspond to relatively high OA loadings (~0.5 OA/nm²), formation of a bridging complex between two surface Fe involving one oxygen of the carboxylate and the oxygen of the keto group of OA (Fig. 4b) is likely expected at the

(100)/(001) plane. In addition, the effect of OA surface coverage is investigated for $[OA]_{\text{tot}}$ down to 20 μM (i.e. 0.24 adsorbed OA molecule per nm^2) at $\text{pHc} \approx 6$ and 100 mM NaCl (Fig. S3c). At such loading, OA surface complexation to the (010) plane of akaganéite is expected to be the major sorption mechanism (Fig. 2b, Fig. S2). No change in the coordination mode of OA on akaganéite can be derived from our ATR-FTIR spectra when $[OA]_{\text{tot}}$ varies. We propose the formation of a MB surface complex involving one oxygen of the carboxylate and the oxygen of the keto group of OA. Because both water molecules of the $\equiv\text{Fe}(\text{OH}_2)_2^+$ groups can be released during the complexation of OA, we propose the formation of a six-membered chelate, involving only one Fe (Fig. 4a). Therefore, the proposed structure for the OA complex at the akaganéite (010) is presently based on crystallographic considerations.

At high pHc, we propose the predominance of a HB complex seen through a band at $\sim 1350\text{ cm}^{-1}$. This band is broad and overlaps with the band at 1330 cm^{-1} , especially for $\text{pHc} = 8.02$ and 8.66 . Its disappearance in D_2O (Fig. S4) suggests that it is sensitive to the hydrogen bonding environment of OA, as in the case of goethite. As no significant ionic strength effects were observed, OA could be directly hydrogen bonded to surface (hydr)oxo groups of akaganéite (Fig. 4c, Fig. S3d).

3.3. Surface complexation modeling

According to the aforementioned crystallographic considerations and our spectroscopic investigations, MB surface complexes at the goethite (101)/(001) and (210) planes (Pnma group) are expected to be comparable to those on the (100) and (001) planes of akaganéite. OA forms a complex involving keto group and one oxygen of the carboxylate group, and it can be expressed in the following reaction:

where (+1;0;0) denotes the charge variations at the 0-, 1- and 2-planes ($\Delta z_0; \Delta z_1; \Delta z_2$). Only singly coordinated ($\equiv\text{FeOH}^{-0.5}$) surface sites are considered to be involved in the latter reaction given the propensity for ligand exchange of these sites. In the case of goethite, note that $\equiv\text{FeOH}^{-0.5}$ is used as a building block for different adsorption sites on the (101)/(001) and the (210) planes. Therefore, Eq. 4 does not denote the number of Fe(III) octahedra in a complex, but only the number of $\equiv\text{FeOH}^{-0.5}$ sites that may or may not be of the same Fe(III) octahedron.⁴⁷ Our results cannot resolve the number of Fe(III) bound to OA explicitly with the data on hand. In two recent MUSIC modeling studies, CIP¹⁰ and OFX¹⁴ were assumed to bind only to the goethite (210) plane. However, from our data, we cannot exclude that OA binds to the goethite (101)/(001) planes.

Hydrogen-bonded (HB) complexation could form by involving various types of surface sites. For the sake of simplicity, only singly $\equiv\text{FeOH}^{-0.5}$ sites are considered:⁴⁷

Preliminary tests showed that $\log K_{\text{HB}}$ and $\log K_{\text{MB}}$ values did not significantly differ for goethite. Therefore, they were constrained at the same value during the fit. The same strategy was applied for akaganéite. According to preliminary tests, it appeared unnecessary to include an additional charge distribution (CD)²⁸ term, unlike previous studies dedicated to CIP¹⁰ and OFX¹⁴ that can bind to mineral surfaces under the zwitterion form.

Finally, to account for OA binding on the Fe sites of the (010) akaganéite plane, an additional reaction was used, which we here describe as a geminal hydroxo/aquo displacement reaction:

370

371 In all of our modeling attempts, we neglect Cl^- surface complexation to the akaganéite (010)
 372 plane highlighted in earlier work^{25,37,38} as our data reveal no ionic strength dependence on OA
 373 sorption. Cl^- surface complexation to the akaganéite (010) plane may be too weak to compete
 374 significantly with OA under the present conditions.

375 Surface complexation reactions and parameters for OA are reported in Table 1.
 376 Results of the fit with PhreePlot are shown in Figure 1 for pH-edges, and in Figure 2b for
 377 sorption isotherms. Our model almost predicts $[\text{NaCl}]$ independent OA loading (Fig. 1a,b).
 378 Loading effects and additional apparent OA uptake on goethite (Fig. 1c), actually due to the
 379 bulk precipitation of OA are also well-predicted. Speciation calculations show that about 50%
 380 of 50 μM of OA precipitate in contact with goethite at $\text{pHc} < 6.5$, in agreement with
 381 spectroscopic observations for samples prepared with high $[\text{OA}]_{\text{tot}}$ and $\text{pHc} < 7$ (Fig. S3a). In
 382 addition, the fraction of hydrogen-bonded OA at the goethite (101)/(001) planes or akaganéite
 383 (100)/(001) planes increases with pHc (Fig. S5), consistently with our FTIR data.
 384 Discrepancies between experimental and modeling results are observed for akaganéite at high
 385 loadings ($[\text{OA}]_{\text{tot}} > 20 \mu\text{M}$; Fig. 1d, Fig. 2a), probably due to steric hindrance and/or
 386 electrostatic effects at high ligand loadings. Nevertheless, the model accurately predicts OA
 387 sorption to akaganéite at environmentally relevant concentrations, i.e. lower OA loadings
 388 ($[\text{OA}]_{\text{tot}} < 20 \mu\text{M}$; Fig. 1b, Fig. 2a).

389 The formation of a strong complex at the akaganéite (010) plane significantly
 390 contributes to the overall uptake, even when the surface is almost saturated with OA. Geminal
 391 sites ($\equiv\text{Fe}(\text{OH}_2)_2^+$) at this plane could be especially reactive for metal-bonded binding, as they
 392 facilitate a mononuclear six-membered chelate complex via the displacement of two
 393 hydroxo/aquo groups at the equatorial plane of a single Fe octahedron. This form of

interaction is generally excluded on the other dominant, yet idealized, (001)/(101)/(210) planes of goethite and (100)/(001) of akaganéite where bridging metal-bonded and hydrogen-bonded complexes are expected to be favored. The contribution of this complex is illustrated e.g. in Figure 1d for $[OA]_{\text{tot}}=50 \mu\text{M}$, where $\equiv\text{Fe}(\text{OA})^0$ represents about 20% of OA uptake. The model predicts that the latter complex prevails at low loadings (Fig. 2b) and high pH. In our experimental conditions, for $[OA]_{\text{tot}}=10 \mu\text{M}$, $[\text{NaCl}]=100 \text{ mM}$ and $\text{pH}=8$, $\equiv\text{Fe}(\text{OA})^0$ accounts for ~65% of OA uptake on akaganéite (Fig. 1b).

4. Environmental implications

Our combined kinetic, liquid chromatographic and ATR-FTIR investigations showed that only adsorption, and not oxidation, controls OA removal from aqueous suspensions of goethite and akaganéite. We demonstrated that OA binds to goethite or akaganéite surfaces through MB bidentate complex formation at circumneutral to low pH and through a strong HB complex at high pH. The dominant MB complex involved direct binding between two surface Fe with keto and carboxylate oxygens of OA, predominantly in a bridging-type complex. This complexation mode was confirmed across a wide range of salinity (10 to 1000 mM NaCl) including seawater-like one (i.e. 599 mM). OA sorption to akaganéite is greater than to goethite, a result that can be attributed to the higher isoelectric point of this mineral, and also the formation of a strong complex at the akaganéite (010) plane. The contribution of this strong binding with geminal sites ($\equiv\text{Fe}(\text{OH}_2)_2^+$) of the akaganéite (010) plane in the overall sorption extent are strongly dependent on OA concentration. This sorption behavior may explain the strong retention of OA observed in sediments containing iron mineral surfaces,¹⁵ but also to mobile FeOOH particles or colloids present in freshwater and marine environments. This may also explain the slow transformation under dark conditions or high persistence of OA in pond waters⁴⁹. These results may therefore have strong implications for

the understanding of both colloidal and aqueous transport of antibacterial agents used in aquaculture.

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Supporting Information Available

Details of the synthesis and characterization of goethite and akaganéite particles, surface complexation modeling; additional results for kinetic sorption/desorption, OA sorption isotherms and ATR-FTIR spectroscopy in H₂O or D₂O; tables with the whole experimental datasets. This information is available free of charge via the Internet at <http://pubs.acs.org/>.

References

1. Le, T. X.; Muneke, Y. Residues of selected antibiotics in water and mud from shrimp ponds in mangrove areas in Vietnam. *Mar. Pollut. Bull.*, **2004**, 49, 922–929.
2. Samuelsen, O.B. Pharmacokinetics of quinolones in fish: a review. *Aquaculture*, **2006**, 255, 55–75.
3. Samuelsen, O.B.; Ervik, A.; Pursell, L.; Smith, P. Single-dose pharmacokinetic study of oxolinic acid and vetoquinol, an oxolinic acid ester, in Atlantic salmon (*Salmo salar*) held in seawater and in vitro antibacterial activity against *Aeromonas salmonicida*. *Aquaculture*, **2000**, 187, 213–224.
4. Rigos, G.; Nengas, I.; Alexis, M.; Troisi, G.M. Potential drug (oxytetracycline and oxolinic acid) pollution from Mediterranean sparid fish farms. *Aquat. Toxicol.*, **2004**, 69, 281–288.
5. Tamtam, F.; Mercier, F.; Le Bot, B.; Eurin, J.; Tuc Dinh, Q.; Clement, M.; Chevreuil, M. Occurrence and fate of antibiotics in the Seine River in various hydrological conditions. *Sci. Total Environ.*, **2008**, 393, 84–95.
6. Gu, C.; Karthikeyan, K.G. Sorption of the Antimicrobial Ciprofloxacin To Aluminum and Iron Hydrated Oxides. *Environ. Sci. Technol.*, **2005**, 39 (23), 9166–9173.
7. Trivedi, P.; Vasudevan, D. Spectroscopic Investigation of Ciprofloxacin Speciation at the Goethite–Water Interface. *Environ. Sci. Technol.*, **2007**, 41 (9), 3153–3158.

- 452 8. Pei, Z.-G.; Shan, X.-Q.; Kong, J.-J.; Wen, B.; Owens, G. Coadsorption of ciprofloxacin
453 and Cu(II) on montmorillonite and kaolinite as affected by solution pH, *Environ. Sci.*
454 *Technol.*, **2010**, 44 (3), 915-920.
- 455 9. Rakshit, S.; Sarkar, D.; Elzinga, E.J.; Punamiya, P.; Datta, R. Mechanisms of
456 ciprofloxacin removal by nano-sized magnetite. *J. Hazard. Mater.*, **2013**, 246-247,
457 221-226.
- 458 10. Gu, X.; Tan, Y.; Tong, F.; Gu, C. Surface complexation modeling of coadsorption of
459 antibiotic ciprofloxacin and Cu(II) and onto goethite surfaces. *Chem. Eng. J.*, **2015**, 269,
460 113-120.
- 461 11. Martin, S.; Shchukarev, A.; Hanna, K.; Boily, J.-F. Kinetics and Mechanisms of
462 Ciprofloxacin Oxidation on Hematite Surfaces. *Environ. Sci. Technol.*, **2015**, 49 (20),
463 12197-12205.
- 464 12. Goyne, K. W.; Chorover, J.; Kubicki, J. D.; Zimmerman, A. R.; Brantley, S. L. Sorption
465 of the antibiotic ofloxacin to mesoporous and nonporous alumina and silica. *J. Coll. Int.*
466 *Sci.*, **2005**, 283, 160-170.
- 467 13. Paul, T.; Machesky, M.; Strathmann, T. Surface Complexation of the Zwitterionic
468 Fluoroquinolone Antibiotic Ofloxacin to Nano-Anatase TiO₂ Photocatalyst Surfaces.
469 *Environ. Sci. Technol.*, **2012**, 46 (21), 11896-11904.
- 470 14. Paul, T.; Liu, J.; Machesky, M.L.; Strathmann, T. Adsorption of zwitterionic
471 fluoroquinolone antibacterials to goethite: A charge distribution-multisite complexation
472 model. *J. Coll. Int. Sci.*, **2014**, 428, 63-72.
- 473 15. Pouliquen, H.; Le Bris, H. Sorption of oxolinic acid and oxytetracycline to marine
474 sediments. *Chemosphere*, **1996**, 33 (5), 801-815.
- 475 16. Burns, R. G.; Burns, V. M. *Authigenic oxides. p. 875-914. In The Oceanic Lithosphere,*
476 *Vol. 7, ed. Emiliani C., John Wiley & Sons, New York. 1981.*
- 477 17. Font, E.; Abrajevitch, A. Paleoenvironmental signature of the Deccan Phase-2
478 eruptions. *Frontiers in Earth Sciences*, **2014**, 2(23), 1-5.
- 479 18. Selwyn, L. S.; Sirois, P. J.; Argyropoulos, V. The corrosion of excavated archaeological
480 iron with details on weeping and akaganéite. *Stud. Conserv.*, **1999**, 44, 217-232.
- 481 19. Gismelseed, A.; Elzain, M.; Yousif, A.; Al Rawas, A.; Al-Omari, I. A.; Widatallah, H.;
482 Rais, A. Identification of corrosion products due to seawater and fresh water. *Hyperfine*
483 *Interact.*, **2004**, 156, 487-492.
- 484 20. Remazeilles, C.; Refait, P. On the formation of beta-FeOOH (akaganéite) in chloride-
485 containing environments. *Corros. Sci.*, **2007**, 49, 844-857.
- 486 21. Bibi, I.; Singh, B.; Silvester, E. Akaganéite (β -FeOOH) precipitation in inland acid
487 sulfate soils of south-western New South Wales (NSW), Australia. *Geochim.*
488 *Cosmochim. Acta*, **2011**, 75, 6429-6438.
- 489 22. Cornell, R. M.; Schwertmann, U. *The iron oxides : structure, properties, reactions,*
490 *occurrences, and uses*; Wiley-VCH: Weinheim, **2003**.
- 491 23. van der Zee, C.; Roberts, D.; Rancourt, D. G.; Slomp, C. P. Nanogoethite is the
492 dominant reactive oxyhydroxide phase in lake and marine sediments. *Geology*, **2003**, 31
493 (11), 993-996.

24. Gaboriaud, F.; Ehrhardt, J. J. Effects of different crystal faces on surface charge of colloidal goethite (α -FeOOH) particles: an experimental and modeling study. *Geochim. Cosmochim. Acta*, **2003**, 67, 967-983.
25. Kozin, P. A.; Boily J.-F. Proton Binding and Ion Exchange at the Akaganéite/Water Interface. *J. Phys. Chem. C*, **2013**, 117, 6409–6419.
26. Tolls, J. Sorption of veterinary pharmaceuticals in soils: a review. *Environ. Sci. Technol.*, **2001**, 35 (17), 3397-3406.
27. Altmaier, M.; Metz, V.; Neck, V.; Müller, R.; Fanghänel, Th. Solid-liquid equilibria of $\text{Mg}(\text{OH})_{2(\text{cr})}$ and $\text{Mg}_2(\text{OH})_3\text{Cl}\cdot 4\text{H}_2\text{O}_{(\text{cr})}$ in the system Mg-Na-H-OH-Cl- H_2O at 25 °C. *Geochim. Cosmochim. Acta*, **2003**, 67, 3595-3601.
28. Hiemstra, T.; van Riemsdijk, W. H. A surface structural approach to ion adsorption: the charge distribution (CD) model. *J. Colloid Interf. Sci.*, **1996**, 179, 488-508.
29. Parkhurst, D. L.; Appelo, C. A. J. *User's guide to PHREEQC (Version 2) – a computer program for speciation, batch reaction, one-dimensional transport and inverse geochemical calculation*; Denver, Colorado, **1999**; p 312.
30. Kinniburgh, D. G.; Cooper, D. M. *PhreePlot: Creating graphical output with PHREEQC*. <http://www.phreeplot.org>. **2009**.
31. Powell, M. J. D. A method for minimizing a sum of squares of non-linear functions without calculating derivatives. *The Computer Journal*, **1965**, 7, 303-307.
32. Venema, P.; Hiemstra, T.; van Riemsdijk, W. H. Multisite Adsorption of Cadmium on Goethite. *J. Coll. Int. Sci.*, **1996**, 183, 515-527.
33. Lützenkirchen, J.; Marsac, R.; Kulik, D. A.; Payne, T. E.; Xue, Z.; Orsetti, S.; Haderlein, S. B. Treatment of multi-dentate surface complexes and diffuse layer implementation in various speciation codes. *Appl. Geochem.*, **2015**, 55, 128-137.
34. Johnson, B. B.; Sjöberg, S.; Persson, P. Surface Complexation of Mellitic Acid to Goethite: An Attenuated Total Reflection Fourier Transform Infrared Study. *Langmuir*, **2004**, 20, 823-828.
35. Norén, K.; Persson, P. Adsorption of monocarboxylates at the water/goethite interface: The importance of hydrogen bonding. *Geochim. Cosmochim. Acta*, **2007**, 71, 5717-5730.
36. Boily, J.-F.; Lützenkirchen, J.; Blamès, O.; Beattie, J.; Sjöberg, S. Modeling proton binding at the goethite (α -FeOOH)–water Interface. *Colloids Surf. A*, **2001**, 179, 11–27.
37. Song X.; Boily, J.-F. Surface Hydroxyl Identity and Reactivity in Akaganéite *J. Phys. Chem. C*, **2011**, 115, 17036-17045.
38. Song X.; Boily, J.-F. Competitive ligand exchange on akaganéite surfaces enriches bulk chloride loadings. *J. Coll. Int. Sci.*, **2012**, 376, 331-333.
39. IUPAC (2001) *IUPAC Stability Constants Database. Version 5.4, IUPAC and Academic Software* (acadsoft@bcs.org.uk).
40. Neugebauer, U.; Szeghalmi, A.; Schmitt, M.; Kiefer, W.; Popp, J.; Holzgrabe, U. Vibrational spectroscopic characterization of fluoroquinolones. *Spectrochim. Acta, Part A*, **2005**, 61, 1505-1517.

- 535 41. Tarushi, A.; Christofis, P.; Psomas, G. Synthesis, characterization and interaction with
536 DNA of mononuclear metal complexes with oxolinic acid. *Polyhedron*, **2007**, 26, 3963-
537 3972.
- 538 42. Nakamoto, K. *Infrared and Raman Spectra of Inorganic and Coordination Compounds.*
539 *Part B: Applications in Coordination, Organometallic, and Biorganic Chemistry*, 5th
540 ed.; Wiley-Interscience: New York, **1997**.
- 541 43. Yost, E. C.; Tejedor-Tejedor, M. I.; Anderson, M. A. In Situ CIR-FTIR
542 Characterization of Salicylate Complexes at the Goethite/Aqueous Solution Interface.
543 *Environ. Sci. Technol.*, **1990**, 24, 822–828.
- 544 44. Cooper, E. M.; Vasudevan, D. Hydroxynaphthoic acid isomer sorption on goethite. *J.*
545 *Col. Int. Sci.*, **2009**, 333, 85–96.
- 546 45. Rusch, B.; Hanna, K.; Humbert, B. Sorption and Transport of Salicylate in a Porous
547 Heterogeneous Medium of Silica Quartz and Goethite. *Environ. Sci. Technol.*, **2010**, 44,
548 2447–2453.
- 549 46. Hanna, K.; Boily, J.-F. Sorption of Two Naphthoic Acids to Goethite Surface under
550 Flow through Conditions. *Environ. Sci. Technol.*, **2010**, 44, 8863–8869.
- 551 47. Boily, J.-F.; Persson, P.; Sjöberg, S. Benzenecarboxylate surface complexation at the
552 goethite (α -FeOOH)/water interface: II. Linking IR spectroscopic observations to
553 mechanistic surface complexation models for phthalate, trimellitate, and pyromellitate.
554 *Geochim. Cosmochim. Acta*, **2000**, 64 (20), 3453-3470.
- 555 48. Yates, J. T. Jr.; Madey, T. E. *Vibrational Spectroscopy of Molecules on Surfaces*.
556 Plenum Press, **1987**.
- 557 49. Lai, H.-T.; Chien, Y.-H. ; Lin, J.-S. Long-term transformation of oxolinic acid in water
558 from an eel pond. *Aquaculture*, **2008**, 275 (1-4), 96-101.
- 559 50. Psomas, G.; Tarushi, A.; Efthimiadou, E. K.; Sanakis, Y.; Raptopoulou, C. P.; Katsaros,
560 N. Synthesis, structure and biological activity of copper(II) complexes with oxolinic
561 acid. *J. Inorg. Biochem.*, **2006**, 100, 1764-1773.

562 **Table 1. Surface Complexation Model.**

Goethite ^a	log K	Δz ₀	Δz ₁	Δz ₂	(101)	(001)	(210)
					(Site/nm ²)		
≡Fe ₃ O ^{-0.5} + H ⁺ ⇌ ≡Fe ₃ OH ^{+0.5}	9.1	+1	0	0	3.03	3.34	-
≡Fe ₃ O ^{-0.5} + H ⁺ + Cl ⁻ ⇌ ≡Fe ₃ OH ₂ ^{+0.5} ...Cl ⁻	8.1	+1	0	-1			
≡Fe ₃ O ^{-0.5} + Na ⁺ ⇌ ≡Fe ₃ OH ^{+0.5} ...Na ⁺	-1	0	0	+1			
≡FeOH ^{-0.5} + H ⁺ ⇌ ≡FeOH ₂ ^{+0.5}	9.1	+1	0	0	3.03	3.34	7.4
≡FeOH ^{-0.5} + H ⁺ + Cl ⁻ ⇌ ≡FeOH ₂ ^{+0.5} ...Cl ⁻	8.1	+1	0	-1			
≡FeOH ^{-0.5} + Na ⁺ ⇌ ≡FeOH ^{+0.5} ...Na ⁺	-1	0	0	+1			
2 H ⁺ + 2 ≡FeOH ^{-0.5} + OA ⁻ ⇌ (≡FeOH ₂) ₂ ⁺ ... (OA) ⁻	20.3±0.1 ^c	+2	-1	0			
2 H ⁺ + 2 ≡FeOH ^{-0.5} + OA ⁻ ⇌ (≡Fe) ₂ (OA) ⁰ + 2 H ₂ O	20.3±0.1 ^c	+1	0	0			

Akaganéite ^b	log K	Δz ₀	Δz ₁	Δz ₂	(100)/(001)	(010)
					(Site/nm ²)	
≡FeO ^{-1.5} + H ⁺ ⇌ ≡FeOH ^{-0.5}	20.0	+1	0	0	3.09	-
≡FeO ^{-1.5} + 2 H ⁺ ⇌ ≡FeOH ₂ ^{+0.5}	28.2	+2	0	0		
≡FeO ^{-1.5} + 2 H ⁺ + Cl ⁻ ⇌ ≡FeOH ₂ ^{+0.5} ...Cl ⁻	29.2	+2	0	-1		
≡FeO ^{-1.5} + H ⁺ + Na ⁺ ⇌ ≡FeOH ^{-0.5} ...Na ⁺	21.0	+1	0	+1		
2 H ⁺ + 2 ≡FeOH ^{-0.5} + OA ⁻ ⇌ (≡FeOH ₂) ₂ ⁺ ... (OA) ⁻	21.5±0.1 ^c	+2	-1	0		
2 H ⁺ + 2 ≡FeOH ^{-0.5} + OA ⁻ ⇌ (≡Fe) ₂ (OA) ⁰ + 2 H ₂ O	21.5±0.1 ^c	+1	0	0		
≡Fe ₃ O _I ^{-0.5} + H ⁺ ⇌ ≡Fe ₃ O _I H ^{+0.5}	11.7	+1	0	0	6.18	-
≡Fe ₃ O _{II} ^{-0.5} + H ⁺ ⇌ ≡Fe ₃ O _{II} H ^{+0.5}	2.6	+1	0	0	3.09	-
≡Fe ₂ O ⁻¹ + H ⁺ ⇌ ≡Fe ₂ OH ⁰	10.8	+1	0	0	3.09	7.06
≡Fe ₂ O ⁻¹ + 2 H ⁺ ⇌ ≡Fe ₂ OH ₂ ⁺	17.9	+2	0	0		
≡Fe(OH ₂)(OH) ⁰ + H ⁺ ⇌ ≡Fe(OH ₂) ₂ ⁺	8.2	+1	0	0	-	3.53
H ⁺ + ≡Fe(OH ₂)(OH) ⁰ + OA ⁻ ⇌ ≡Fe(OA) ⁰ + 2 H ₂ O	13.9±0.1	0	0	0		

^aTPM with $C_1=2.3 \text{ F/m}^2$, $C_2=1.07 \text{ F/m}^2$, surface area: 95 m²/g; 63% of (101), 27% of (001) and 10% of (210). The HB surface complex is considered to form only at the (101) and (001) planes while the MB occurs at all planes.

^bTPM with $C_1=2.3 \text{ F/m}^2$, $C_2=1.6 \text{ F/m}^2$, surface area: 239 m²/g; 95% of (100)/(001) and 5% of (010).

^cConstant and uncertainty obtained with Phreeplot, corrected according to Venema et al.³².

Figure captions

Figure 1. OA removal from solution (molecule / nm²) for [OA]_{tot}=10 μM on 50 m²/L goethite (a) and akaganéite (b) versus pHc (= -log [H⁺]) at different NaCl concentrations. OA removal from solution (molecule / nm²) for [OA]_{tot}=50 μM on 50 m²/L goethite (c) and akaganéite (d) in 100 mM NaCl versus pH (= -log [H⁺]). The corresponding percentage of OA removal at the plateau is also given. Lines are modeling results, where OA speciation on the solid phase is shown for [NaCl]=100 mM (legend in a/c and b/d is the same for the model). MB and HB refer to metal- and hydrogen-bonded OA, respectively.

Figure 2. (a) Experimental and calculated OA solubility in 100 mM NaCl versus pHc. The equilibrium OA concentration ([OA]_{aq}) in contact with goethite and akaganéite for [OA]_{tot}=10 and 50 μM in 100 mM NaCl are also plotted with modeling results. (b) OA-goethite and OA-akaganéite sorption isotherms at pHc=6.1±0.1 and 100 mM NaCl. The results of the models for goethite and akaganéite and OA speciation at the akaganéite surface are also shown. OA solubility limit at pHc=6.1 is shown as a vertical dashed line.

Figure 3. ATR-FTIR spectra of OA adsorbed to goethite (a) and akaganéite (b) for various pHc (-log [H⁺]) in 100 mM NaCl for [OA]_{tot}=50 μM (except for goethite at pHc=5.20 and 6.39, where [OA]_{tot}=20 μM). ATR-FTIR spectra of dissolved OA in 1 M NaOH (OA_(aq)) and protonated OA under its solid form (OA_(s)) are shown as references.

Figure 4. Possible surface complexes of OA at goethite and akaganéite surfaces. The electronic structure of OA is proposed according to previous studies on metal-OA compounds.^{41,50}

Figure 1

Figure 2

Figure 3

Figure 4

TOC/abstract